

A life-cycle analysis of deep enhanced geothermal systems – the case studies of Reykjanes, Iceland and Vendenheim, France

Abstract

The climate impacts of deep enhanced geothermal systems (EGS) have been understudied in the academic literature. Using life-cycle analysis (LCA) conducted in accordance with ISO 14040 and ISO 14044 standards, this paper explores the climate change impacts of two deep EGS. The first study was in Reykjanes, Iceland, where a single well, IDDP-2/DEEPEGS, was drilled to a depth of 4.6 km for the purposes of additional electricity production from an existing power plant. The second study involved two wells with side-tracks (depth > 5,000 m), drilled to serve a new heat and power co-generation plant located on an old oil field site in Vendenheim, France. Climate change impacts for the sites were estimated in the range 1.6–17.4 gCO₂e/kWh and 6.9–13.9 gCO₂e/kWh for Reykjanes and Vendenheim, respectively. Although the EGS projects are very different, both outcomes are low when compared to non-renewable alternatives and akin to best-in-class renewable alternatives. The main impact at the Reykjanes demonstration site were the greenhouse gas emissions released from the borehole, an effect that could be avoided by carbon capture and storage/mineralisation/utilisation. In the case of Vendenheim, further reductions in emissions could be achieved via more extensive adoption of circular economy principles in design and procurement.

Keywords: life-cycle assessment; deep geothermal energy; environmental impacts; hotspots; greenhouse gas emissions

48 **1. Introduction**

49
50 1.1 Utilisation of geothermal energy resources and enhanced geothermal systems
51
52 For well over 100 years, geothermal resources have provided base-load electricity around the
53 world (Tomasini-Montegro et al., 2017). Total worldwide installed capacity from geothermal
54 technologies was 15.9 GW in 2019 (Huttrer, 2020), with likely future growth in the utilisation
55 of conventional and enhanced geothermal systems (EGS) (Aghahosseini et al., 2020; Pan et al.,
56 2020). The difference between a conventional geothermal system and EGS is that in principle
57 an EGS could theoretically harness the heat in every place on the earth, by enhancing the
58 properties of the geothermal system so that it can be utilised similarly to conventional systems
59 (Olasolo et al., 2016). EGS reservoirs are deep enough to reach a high-temperature rock,
60 however, usually without any fluid, and need to have an injection well/wells to pump water
61 down to it (Xu et al., 2015). Nevertheless, the rock in the reservoir is either impermeable or has
62 low permeability, preventing any heat extraction (Shao et al., 2015). As a result, the reservoir
63 needs to be stimulated and enhanced for the water to be able to travel through it and heat up
64 before coming back up to the production well and being utilised in a power conversion unit
65 (Huenges and Ledru, 2011; Olasolo et al., 2016).

66
67 Since the early 1970s, there have been numerous EGS projects in Europe, the USA, Australia
68 and Japan, which have sought to utilise the energy stored in the fluid-impaired geothermal
69 reservoir at that site (Olasolo et al., 2016). The initial results were that deep high-temperature
70 wells could be drilled and completed in the hard, abrasive rock at these EGS sites. Furthermore,
71 a reservoir could be stimulated to a large enough extent to sustain energy extraction over a long
72 period of time (Tomasini-Montegro et al., 2017). Heat in a reservoir as deep as 5,000m below
73 the surface can be accessed and extracted (Tester et al., 2006). Thus, in finding the most
74 prominent areas for an EGS on a European scale, a map of temperature extrapolated at 5 km
75 depth, conducted by Genter et al. (2003) and modified by the Economic Interest European
76 Group (EIEG), shows the potential feasibility of EGS as a climate change mitigation technology
77 in Europe. The most promising areas are thought to be the Pannonian Basin, French Massif
78 Central, Rhine Graben Basin, the Campidano Graben, Tuscany and a narrow graben in central
79 Greece. Other areas with potential include the Limagne Basin, the Rhine Graben Borders, the
80 Rhone Valley and the Provence, the North German Basin, Urach, Catalonia, the northern part
81 of Bulgaria, and Cornwall in the United Kingdom.

82
83 1.2 Environmental impacts of geothermal energy and life-cycle analysis (LCA)

84
85 Given the likely expansion in production, increased attention has been directed towards the
86 environmental impacts of the geothermal power sector, particularly in connection with the
87 harnessing of high-temperature fields for electricity generation. Several studies have described
88 the environmental impacts of geothermal power projects and their effects on societal well-being
89 (Ármannsson & Kristmannsdóttir, 1992; Bayer et al. 2013; Shortall et al., 2015; Cook et al.,
90 2020). Bayer et al. (2013) observed that these include direct effects¹, such as land distortion
91 (land use), atmospheric emissions (especially carbon dioxide and hydrogen sulphide), waste
92 heat, solid waste (terrestrial ecotoxicity potential), water consumption (water demand), noise
93 emissions (human toxicity potential), and impacts on biodiversity. However, there also exist a
94 variety of indirect effects specific to the harnessing of geothermal power. These often relate to
95 the materials and energy required over the lifecycle of the power plant (Tomasini-Montegro et
96 al., 2017; Tosti et al., 2020). Therefore, to fully understand the environmental implications of

¹ Examples of associated LCA impact categories stated in brackets.

97 geothermal power generation, a product system approach should be favoured, since this
98 facilitates an assessment which is considerate of the impacts of supply chains over a project's
99 life-cycle. Evaluations of this type can be conducted using Life-Cycle Analysis (LCA), which,
100 when conducted according to well-defined ISO norms and procedures, such as ISO 14040
101 (2006a) and ISO 14044 (2006b), can compile and aggregate all inputs, outputs and
102 environmental impacts of geothermal power generation across a project's lifecycle (Hunkeler
103 et al., 2008; Swarr et al., 2011).

104 In recent years, given recognition of the need to tackle climate change, the many direct and
105 indirect environmental impacts of power generation have assumed increased importance. Even
106 fairly early examples in the academic literature hint at a breadth of focus in LCA studies with
107 respect to the power sector. They include studies on the greenhouse gas emissions of a coal
108 fuelled power plant (Wu et al., 2018) and substituting switch grass for coal in electricity
109 generation (Ney and Schnoor, 2002), an LCA on a natural gas combined-cycle power system
110 (Spath and Mann, 2000), and LCAs focused on wind turbines (Batumbya et al., 2006; Chen et
111 al., 2011). There are also several LCA studies in the academic literature that focus on
112 geothermal power plants (Karlsdóttir et al., 2015; Paulillo et al., 2019; Karlsdóttir et al., 2020).
113 These include a theoretical study on the potential LCA impacts of EGS (Clark et al., 2012), an
114 applied LCA study on geothermal power generation using supercritical steam (Frank, et al.,
115 2012), a review study on the environmental issues of relevance to geothermal LCAs in the
116 context of a Californian plant (Sullivan et al., 2012), an analysis of how to integrate life cycle
117 analysis and energy synthesis for a dry steam geothermal plant in Italy (Buonocore et al., 2015),
118 a comprehensive review of existing LCA studies using different geothermal technologies
119 (Tomasini-Montenegro et al., 2017), a study focused on climate change impacts in relation to a
120 power project in the Upper Rhine Valley (Pratiwi et al., 2018), and a comparison of the life-
121 cycle impacts of a geothermal power venture in Italy compared to wind and solar power (Basosi
122 et al., 2020). As far as the authors are aware, there are no applied LCA studies on EGS.

125 1.3 Aims and case studies

126 Researchers have dedicated an extensive amount of time in modelling, predicting and
127 optimising the hydraulic and thermal aspects of EGS reservoirs (Aghahosseini et al., 2020; Pan
128 et al., 2020). This research, among others, has contributed to making EGS systems technically
129 feasible today. Their utilisation is therefore increasing. However, in contrast to research on
130 conventional geothermal systems, little is known about the environmental impacts of EGS
131 systems, especially those involve deep drilling to depths of 4-6 km. This paper therefore seeks
132 to use LCA explore two research questions:

- 135 • What is the climate change impact of deep EGS and where in the life-cycle do those
136 impacts occur?
- 137 • How can the impacts be potentially mitigated?

138 The paper answers these questions through two deep EGS case studies, Reykjanes, Iceland and
139 Vendenheim, France, both of which are demonstration sites in the European Union-funded
140 Deployment of Deep Enhanced Geothermal Systems for Sustainable Energy Business
141 (DEEPEGS) project. DEEPEGS is an innovative four-year project led by the Icelandic power
142 company, HS Orka, alongside other partners in Iceland, France, Germany, Italy and Norway.
143 The project includes the testing of deep EGS projects in Reykjanes, Iceland and Vendenheim,
144 France. The goal of the project was to demonstrate the feasibility of deep EGS in different
145 locations.

147
148 An existing 2.5 km deep well at Reykjanes, known as IDDP-2, was deepened to 4.6 km, with
149 drilling completed in late 2017, and well stimulation and testing undertaken thereafter. In
150 theory, drilling to unconventional depths enables access to very high temperatures, i.e. 400-
151 500°C in hydrothermal areas and about 200°C in hot, dry rock areas (Friðleifsson et al., 2014;
152 Friðleifsson et al., 2020). The hypothesis of the DEEPEGS project was that the consequence of
153 deep drilling projects would be a considerable reduction in costs and environmental impacts
154 (Peter-Borie et al., 2019; Friðleifsson et al., 2020). Far fewer wells should, in theory, be required
155 to generate the same power output using regular geothermal boreholes and potentially lower
156 gas emissions from boreholes would result, as was demonstrated by Ármansson et al. (2010)
157 in the IDDP-1 project in northern Iceland.

158
159 The Vendenheim dry steam project is led by Fonroche Géothermie and involves the drilling of
160 two deep wells in excess of 5,000 metres depth, with side-tracks starting below depth of 3,500
161 metres to increase connectivity between the VDH1 and WDH2 wells, and subsequent
162 construction of power plant infrastructure on the site of a former fossil fuel plant. The
163 demonstration site at Vendenheim constitutes a traditional EGS, as the heat is the only naturally
164 occurring element in the area, utilised by enhancing the resource-enabling fluid through an
165 injection well to pump down the fluid and induce soft stimulation activity, leading to
166 permeability.

167
168 The two demonstration sites at Reykjanes and Vendenheim were thus very different EGS, as
169 although they were utilising the same fundamental resource, i.e., energy from within the earth
170 geothermal energy, their accessibility and the project characteristics were dissimilar. In
171 Reykjanes, there was already electrical production from the resource categorised as
172 hydrothermal geothermal energy, i.e. there was access to fluid, heat, and permeability in a
173 naturally occurring geological formation at a depth of about 2.5 km. Unlike Reykjanes, where
174 the aim was to provide additional electricity generation from a single well to an established
175 power plant, the Vendenheim project entailed new co-production of heat and power. Therefore,
176 the results from the two studies cannot be directly compared. Instead, they should be considered
177 as two standalone contributions to the nascent academic literature on the climate change
178 impacts of deep EGS over the duration of project life-cycles.

179
180 **1.4 Structure**

181
182 The remainder of this paper is structured as follows. Section 2 details the LCA modelling
183 methodology with respect to the two case studies, including a conceptualisation of their
184 respective processes, description of the operational characteristics of the sites, and detailing of
185 data sources for the purposes of the LCA inventories. Section 3 provides the results and Section
186 4 discusses these in comparison to other energy-generating technologies and opportunities for
187 further mitigation of impacts. Section 5 outlines a short conclusion and considers options for
188 future research.

189
190 **2. Materials and methods**

191
192 **2.1 Overarching method**

193
194 The LCA methodology for these studies was conducted in accordance with the standards
195 outlined by ISO 14040 (2006a) and ISO 14044 (2006b). As described in recent paper by Basosi
196 et al. (2020), there were four chronological stages to the method. These were as follows:

- 197
198 1) Goal and scope definition
199 2) Life-cycle inventory analysis
200 3) Life-cycle impact analysis
201 4) Life-cycle interpretation

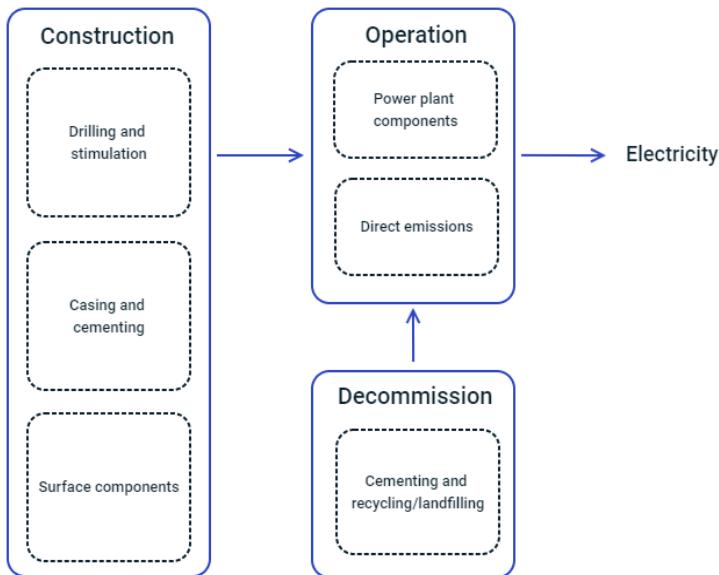
202
203 In the first stage, the goal of the project and its system boundaries were described, quality
204 requisites of data sources scrutinised, and the functional unit of the analysis determined. Stage
205 two involved the collection of input-output data of relevance to the system, with primary data
206 always preferred wherever possible. In the third stage, the climate change impacts of the
207 respective sites were evaluated through the linking of inventory data with specific LCA impact
208 categories and characterisation factors. The Life-Cycle Impact Assessment (LCIA) was
209 conducted in accordance with the Intergovernmental Panel on Climate Change 2013 method
210 (de Bruyn et al., 2018) to find the climate change impact of the demonstration sites. A
211 comparison was then made between the values of the demonstration sites and results using EU-
212 average data for electricity production. As is customary in LCA studies, sensitivity analysis was
213 also conducted during this third stage using Monte Carlo simulation, ensuring that data
214 uncertainty was accounted for. Fourthly, the results of the LCA models were evaluated in the
215 light of opportunities to reduce greenhouse gas emissions across the life-cycles of the projects.

216
217 2.2 Conceptualisation of main operational processes

218
219 2.2.1 Reykjanes

220
221 The expected higher temperatures from drilling the IDDP-2 well could be used to generate
222 superheated steam, increasing electricity production from the site to levels higher than in
223 conventional geothermal power projects (Friðleifsson et al., 2020). In terms of eventual power
224 output, it was originally expected² that this might equate to 25-50 MW from a single well, much
225 more than the typical 4-5 MW from a conventional hydrothermal borehole (Friðleifsson et al.,
226 2020). Figure 1 shows a conceptual drawing of the main processes in the Reykjanes case study
227 as modelled in this work.

² Unfortunately, due to engineering and stimulation complexities, the authors were informed by the DEEPEGS project managers that the most likely power outcome at the end of the project was akin to a conventional geothermal well of 4-5 MW, and the higher expectation of 25-50 MW was desired through further stimulation and other technological innovations, but deemed very unlikely to be attained.



229
230 **Figure 1. A conceptual drawing of the main processes in the Reykjanes case study.**
231

232 The product system at Reykjanes is divided in Figure 1 into construction, operation and
233 decommission stages. The main processes in the construction part were drilling and stimulation
234 of the borehole, casing and cementing the borehole, and the surface components. The
235 operational process was divided into power plant components and direct emissions, and
236 included emissions linked to maintenance. Decommissioning involved a process of disposal,
237 including cementing the borehole and recycling/landfilling the surface and power plant
238 components. Table 1 gives an overview of the main operational characteristics of the Reykjanes
239 case study for two power output scenarios: scenario 1 (most likely) and scenario 2 (desired).
240

241 **Table 1: Main operational characteristics of the Reykjanes project – scenarios 1 and 2.**

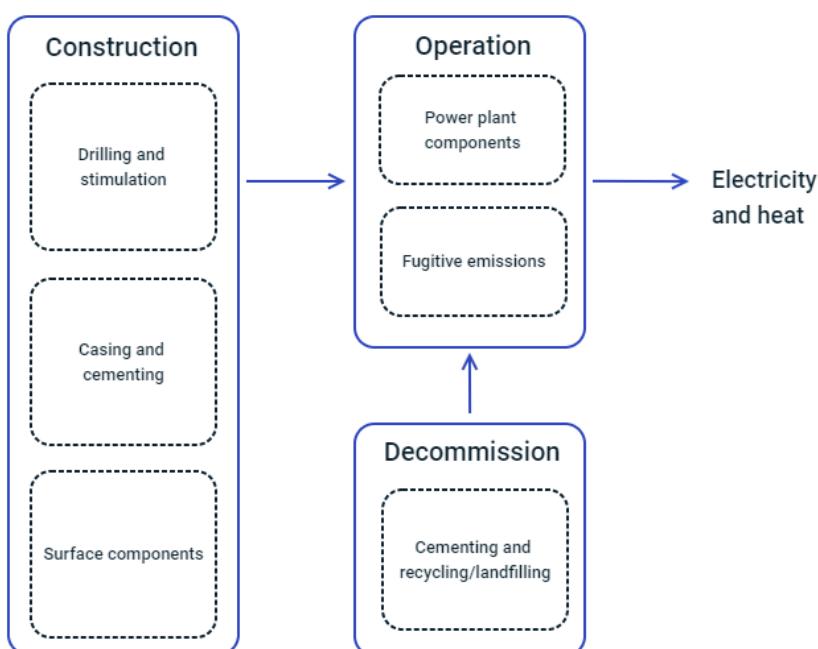
Characteristic	Value (scenario 1)	Value (scenario 2)	Unit
Massflow	50	50	kg/s
Lifetime	30	30	years
Power output	5	30	MW
Steam ratio	18%	100%	
CO ₂	2,328	202	mg/kg steam
CH ₄	0.22	0.02	mg/kg steam
Capacity factor	0.8	0.8	
Annual electricity production	35,040	210,240	MWh
Annual steam production	227,059,200	1,261,440,000	Kg

242 All data on the most likely scenario was provided by HS Orka, the Icelandic operator of the
243 Reykjanes power plant. The most likely scenario assumed that the inflow to the borehole was
244 from the same reservoir as the existing power plant at Reykjanes, and the desired scenario
245 assumed that the inflow is sourced from a deeper reservoir. In the most likely scenario in Table
246 1, the power capacity is around 5 MW, and the gas emissions are 2,328 mgCO₂/kg steam and
247 0.22 mgCH₄/kg steam. The steam ratio in the mass flow is about 18%. In the optimal scenario,
248 based on estimations derived from the preceding IDDP-1 project in northern Iceland, the power
249 capacity is around 30 MW, and the gas emissions are 202 mgCO₂/kg steam and 0.02 mgCH₄/kg
250

251 steam. The steam ratio in the mass flow is 100%. These values were formed from the difference
252 between the “deep” borehole and other boreholes in the area (Ármannsson et al., 2014).

254 2.2.2 Vendenheim

255 The Vendenheim site involved drilling two boreholes with side-tracks for increased
256 connectivity between the wells to a depth of approximately 5 km, where the temperature is
257 200°C. It will be a three-legged doublet system and a combined heat and power plant. The
258 expected electrical efficiency is about 16.5%, resulting in 6-12 MW electricity generated to the
259 grid, with a heat supply capacity factor of approximately 50%. Figure 2 illustrates a conceptual
260 drawing of the main processes in the Vendenheim case study as modelled in this work.



263
264 **Figure 2. A conceptual drawing of the main processes of the Vendenheim case study.**

265 As per the Reykjanes case study, the Vendenheim product system is divided into construction,
266 operation (including emissions from maintenance) and decommission phases. The main
267 processes in the construction part are the same, apart from a higher number of boreholes
268 compared to the Reykjanes case study. The power plant components represent an Organic
269 Rankine Cycle power plant, and the decommissioning part involves cementing the boreholes
270 and recycling/landfilling the surface and power plant components. Table 2 provides an
271 overview of the main operational characteristics of the Vendenheim case study for two power
272 output scenarios: scenario 1 (lower) and scenario 2 (upper).

273 **Table 2. Main operational characteristics of the Vendenheim project – scenarios 1 and 2.**

Characteristic	Value (Sc. 1)	Value (Sc. 2)	Unit
Wells	3	3	Items
Legs	2	2	Items
Temperature at wellhead	200	200	Degrees Celsius
Massflow	80	160	kg/s
Lifetime	30	30	Years
Gross power	40.1	80.3	MW

Electrical efficiency	16.5%	16.5%	
Auxiliary power downhole pump	0.4	0.9	MW
Auxiliary power reinjection pump	0.3	0.6	MW
Power output	5.9	11.8	MW
Heat output	13.4	26.8	MW
Capacity factor electric	0.8	0.8	
Capacity factor heat	0.5	0.5	
Annual electricity production	41,369	82,739	MWh
Annual heat production	58,605	117,210	MWh

276

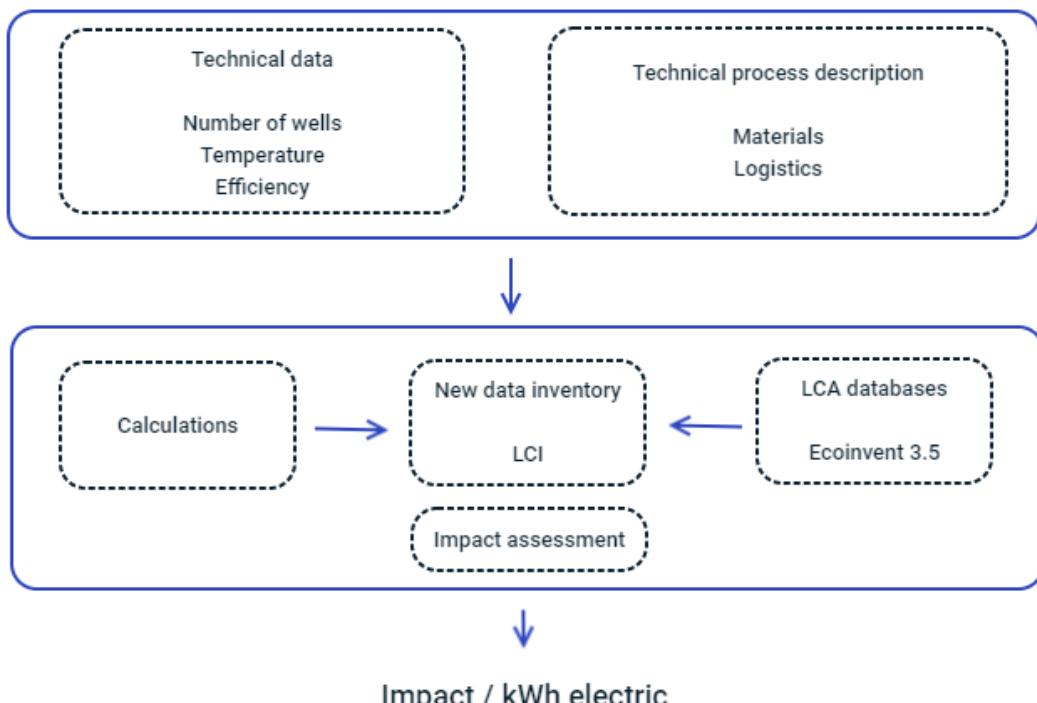
277 The working fluid in the Organic Rankine Cycle was 1233 RZ, which has a low global warming
 278 potential. The main difference between the scenarios concerned the expected flow rate in the
 279 production wells. The production well flow rate is a measure of the enhancing success, i.e. how
 280 well the stimulation activities work to create permeability and induce connection to the injected
 281 fluid from the injection wells. The upper and lower limits were determined based on the findings
 282 of Lacirignola and Blanc (2013).

283

284 2.3 Functional unit and LCA modelling approach

285

286 As the construction of the demonstration site in Reykjanes was (and is) ongoing, the life-cycle
 287 perspective was estimated by modelling the complete product system according to the 1 kWh
 288 electric functional unit. The model then computes results in mean, upper and lower bounds
 289 using Monte Carlo simulation for the two case studies (Metropolis & Ulam, 1949; Mooney,
 290 1997). With respect to the scenarios described in Tables 1 and 2, the differences between them
 291 are mainly productivity and emissions from the borehole as described. This is where most of
 292 the uncertainty is in the operational phase, as these parts of the product system are not completed
 293 yet. Figure 3 is a conceptual drawing of the modelling and simulation approach, from the
 294 gathering of technical data and process descriptions through to impact assessment calculations.



295
296
297 **Figure 3. Conceptual drawing of the modelling and simulation approach.**
298

299 This work relied on the information given by the DEEPEGS partners: HS Orka and Fonroche
300 for the Reykjanes and Vendenheim case studies, respectively. Technical data and technical
301 process descriptions were gathered from those partners, which then formed the foundation of
302 the life-cycle inventory. All scenarios relied on a base model for the life-cycle inventory that
303 was modelled using actual and estimated data inputs based on the construction status of the
304 demonstration sites. The base model represents parts of the life-cycle additional to the
305 operational phase, since that part is more uncertain and depends on the enhancement success,
306 as was described earlier. All data inputs have been given an uncertainty distribution based on
307 data quality using the Pedigree matrix (Ciroth et al., 2013; Ciroth et al., 2016; Muller et al.,
308 2016). The model's likelihood was then assessed using the Monte Carlo simulation.
309

310 2.4 Life-Cycle Inventory

311 This section gives an overview of the main inputs in the life-cycle inventory for the life-cycle
312 assessment. In addition to the data for the main operational characteristics, these inputs are used
313 to find the estimated impact for the two case studies and all scenarios. Tables 3-8 are included
314 in the Appendix and outline the primary material, energy, and transportation processes in the
315 construction, operation and decommission parts of the products' systems.
316

317 The inputs in Tables 3-8 were then normalised into the functional unit, i.e. 1 kWh of electricity
318 produced. The values used to normalise are stated in Tables 1 and 2 for the two cases and all
319 scenarios. These values were divided by the annual electricity production then multiplied by
320 the assumed lifetime of the power projects, which was a duration of 30 years.
321

322 The main differences in Tables 3-8 concern the volume of materials used. These were far greater
323 in the case of Reykjanes, since this project involved the drilling of two new boreholes and a
324 sidetrack. Additionally, gas emissions in the Reykjanes case study are significantly higher than
325

326 the fugitive emissions in the Vendenheim case study, which results in a very different
327 operational impact.

328

329

330 3. Results

331

332 The results are presented separately for the Reykjanes and Vendenheim case studies and their
333 respective power output scenarios. The two power output scenarios for the Reykjanes site were
334 5 MW (lower) and 30 MW (upper); the two power output scenarios (co-generated heat and
335 electricity) for Vendenheim were 40.1 MW and 80.3 MW. Monte Carlo simulation results detail
336 the impacts of the construction phases of the two cases, giving a range for the operational
337 scenarios described in the methods section. The life-cycle impact of greenhouse gas emission
338 hotspots is compared to the carbon intensity of average electricity production in Europe.

339

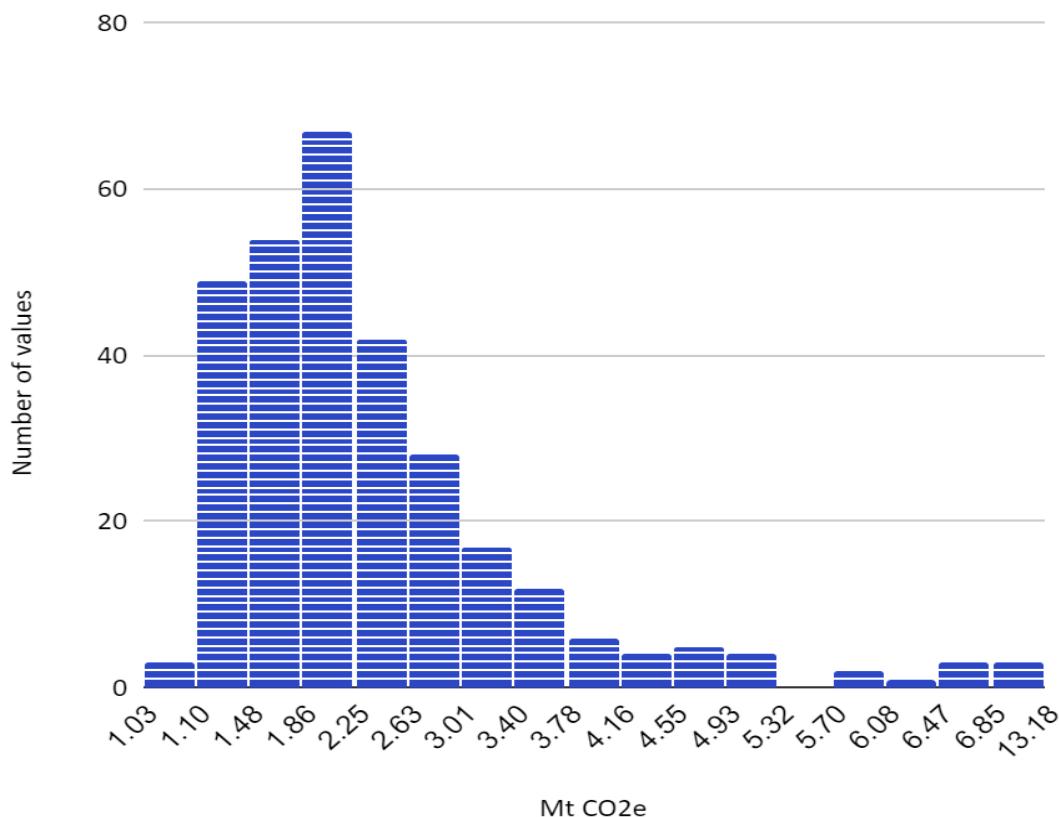
340 3.1 Reykjanes

341

342 The life-cycle impact category hotspots for the Reykjanes case study were split as follows:
343 climate change (45%); particulate matter formation (29%); terrestrial acidification (14%); and
344 human toxicity (8%). For all these impact categories, the impact of average European electricity
345 production is considerably higher than in the Reykjanes case study.

346

347 Figure 4 gives the results of the Monte Carlo simulation for the construction part of the
348 DEEPEGS project at Reykjanes. It includes the borehole construction, stimulation and above-
349 ground equipment.



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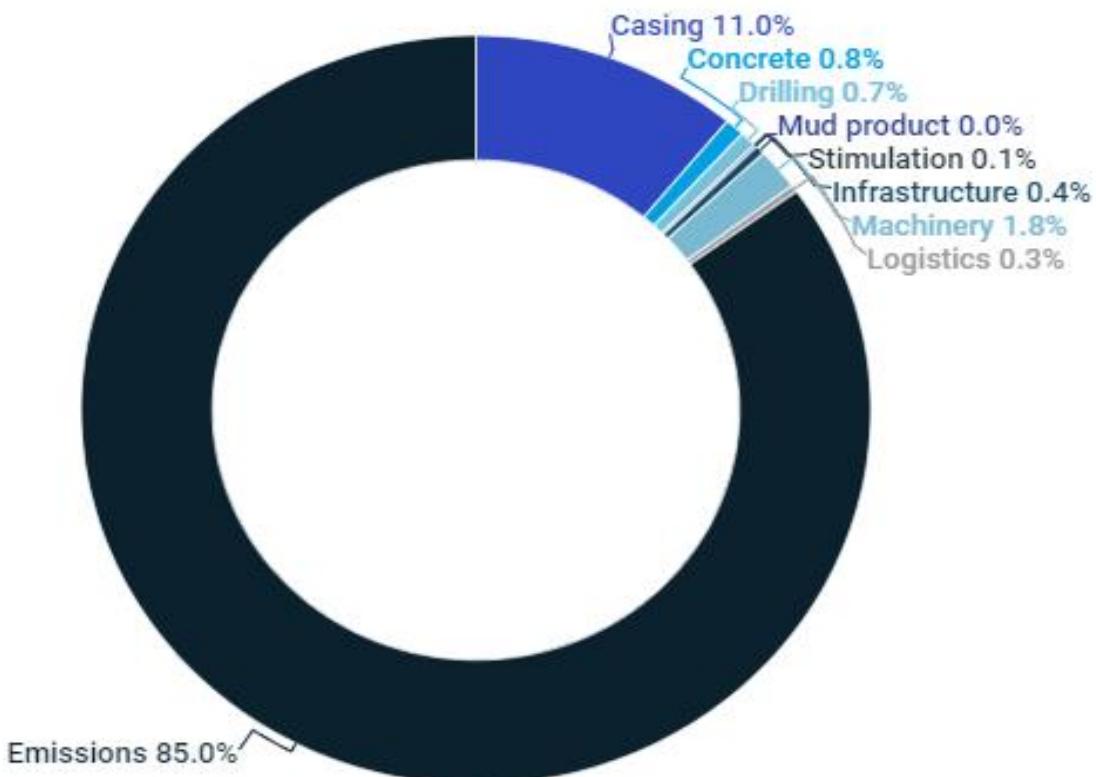
351 **Figure 4. Monte Carlo simulation results of the Reykjanes case study in MtCO₂**
352 **equivalents.**

353

354 The result of the simulation is a mean value of 2.4 gCO₂e/kWh, and a standard deviation of 1.3
355 gCO₂e/kWh. The upper limit based on the 95th percentile is 4.8 gCO₂e/kWh, and the lower limit
356 based on the 5th percentile is 1.2 gCO₂e/kWh. These results were then articulated with respect
357 to the functional unit (1 kWh electricity produced) and combined with the operational impact
358 based on the two scenarios defined in the methods section. The mean result for scenario 1 (most
359 likely) is that the impact of the Reykjanes case study is expected to be 17.4 gCO₂e/kWh, with
360 an upper limit of 19.7 gCO₂e/kWh and a lower limit of gCO₂e/kWh. The results for scenario 2
361 (desired) imply a considerably lower impact – the mean impact is expected to be 1.6
362 gCO₂e/kWh, with an upper limit based on the 95th percentile of 2.0 gCO₂e/kWh and a lower
363 limit based on the 5th percentile of 1.4 gCO₂e/kWh.

364

365 Figures 5 and 6 show the distribution of impacts from different processes within the product
366 system based on scenarios 1 and 2, respectively, which can help to identify opportunities for
367 improvements in sourcing and processing.



368

Figure 5. Distribution of climate change impact in scenario 1 of the Reykjanes case study.

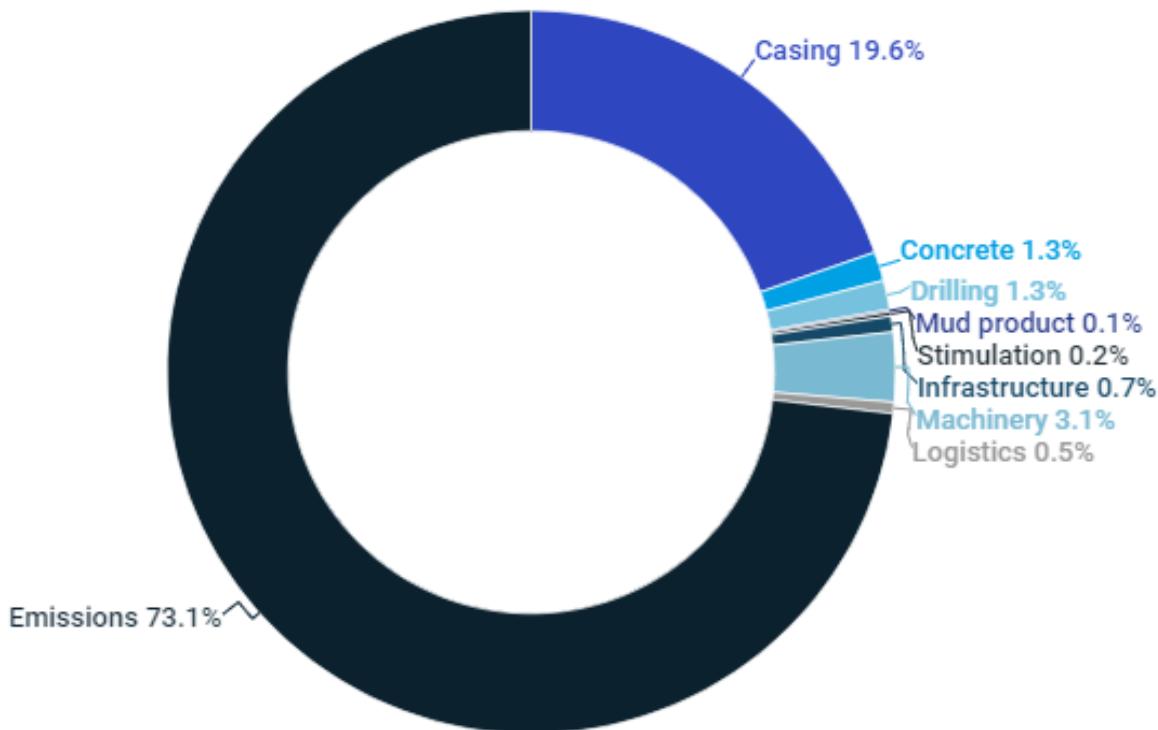
369

370 As Figure 5 above shows, 85% of the climate change impact is from operational emissions i.e.
371 the gas emissions (CO₂) from the borehole which are derived from the geothermal reservoir.
372 Most of the other impacts relate to borehole construction, especially from the steel material
373 (11%) used in the casing.

374

375 Figure 6 below reveals a slight difference in the distribution of climate change impacts between
376 the two scenarios. Operational emissions are relatively less prevalent, and about 20% of the
377 effects are from steel production used in the casing in the borehole. However, when analysing
378 the distribution of impacts from other impact categories for both scenarios, most of the impact
379 is from steel production for the casing i.e. 62% of the particulate matter formation impact and
380

381 64% of the human toxicity impact, except for the terrestrial acidification impact where 64% of
382 the impact is due to operational ammonia emissions from the borehole and only 23% from the
383 steel material used in the casing.



384
385 **Figure 6. Estimated climate change impact in scenario 2 of the Reykjanes demonstration**
386 **site.**

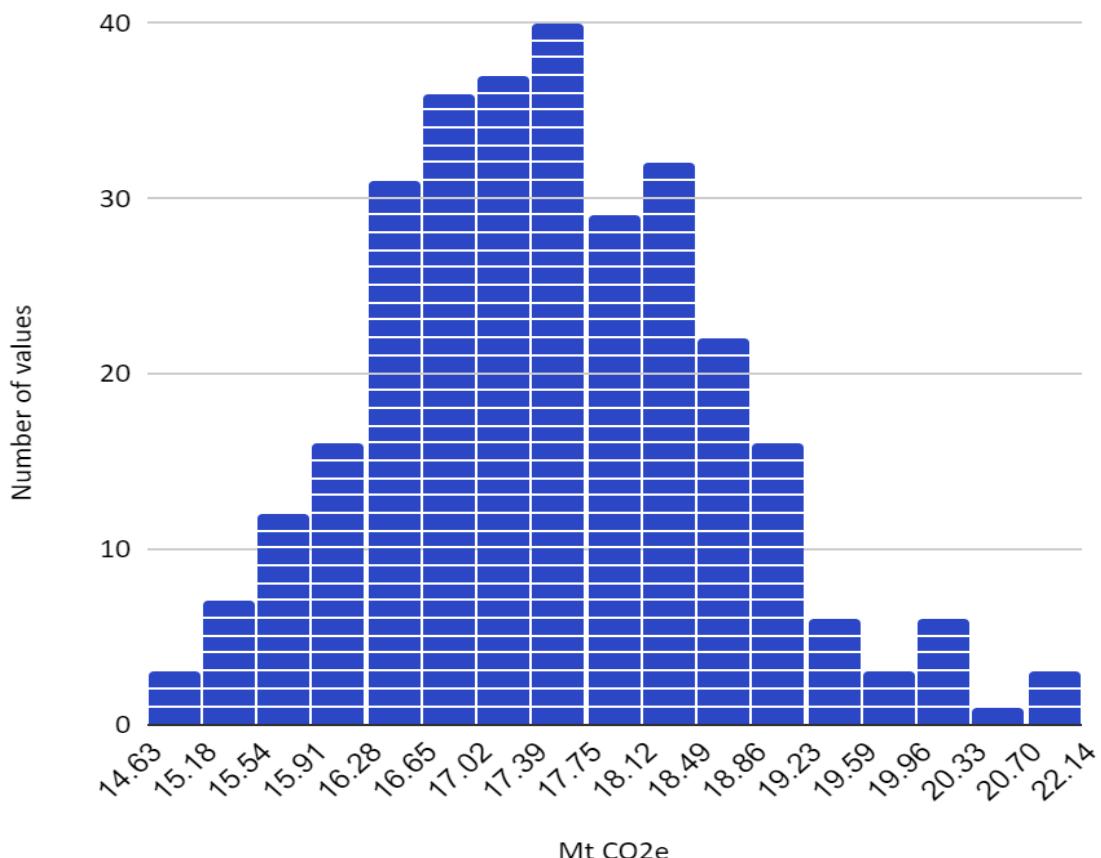
387
388 To summarise the results of the life-cycle perspective for the Reykjanes case study, the most
389 critical impact categories are climate change, particulate matter formation, terrestrial
390 acidification and human toxicity. As this study has been conducted when the Reykjanes
391 demonstration site is still under construction, the impact analysis was conducted using Monte
392 Carlo simulations, and the uncertainty distribution was found using the Pedigree method as
393 described in the methods section. The impact is then estimated over a range for the two
394 scenarios, giving expected results in the range 1.4-19.7 gCO₂e/kWh³.
395
396

397 3.2 Vendenheim

398
399 The life-cycle impact category hotspots for the Vendenheim case study were split as follows:
400 particulate matter formation (30%); human toxicity (25%); climate change (14%); ionising
401 radiation (13%); terrestrial acidification (8%); and terrestrial ecotoxicity (5%). For all these
402 impact categories, the impact of average European production is considerably higher than in
403 the Vendenheim case study.
404

³ Today, based on feedback from the operator, HS Orka, it is considered more likely that the impact when the project is operational will be closer to the 19.7 gCO₂ equivalents, similar to the current situation at Reykjanes power plant as accessing a deeper reservoir has potentially been unsuccessful.

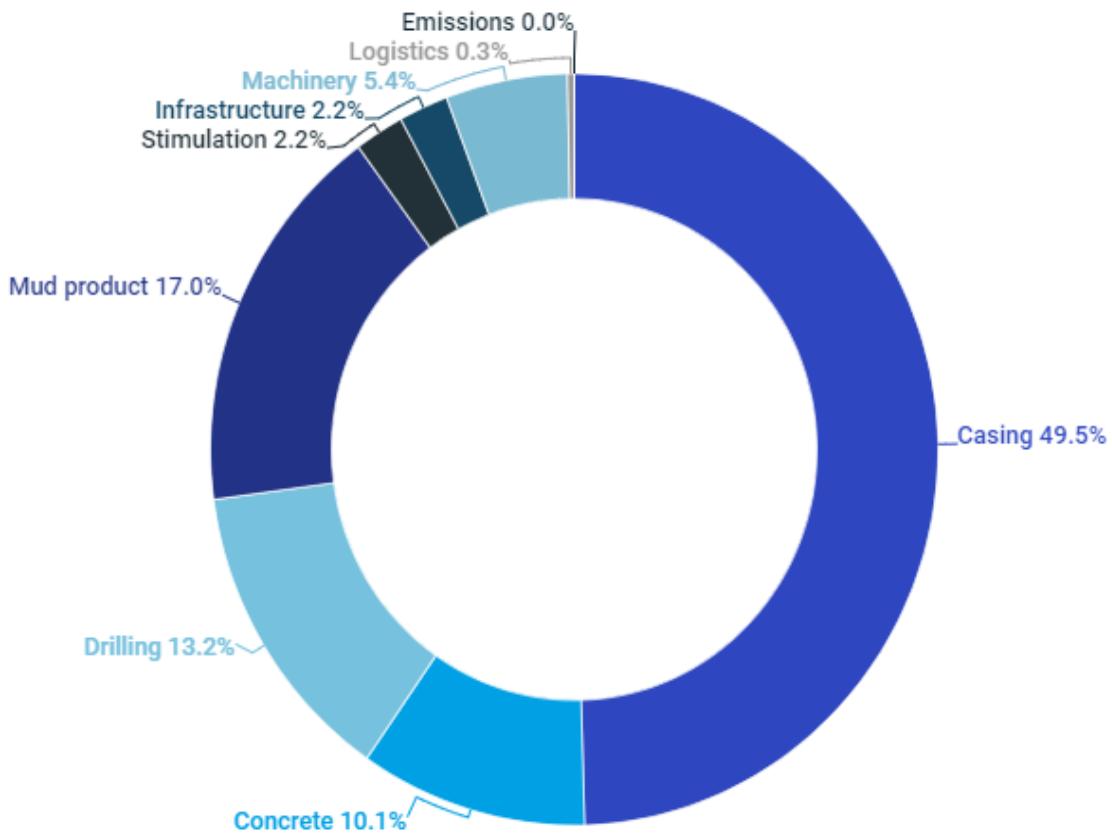
405 Figure 7 depicts the results of the Monte Carlo simulation for the DEEPEGS project at
406 Vendenheim. It includes the borehole construction, stimulation and above-ground equipment.



407
408 **Figure 7. Monte Carlo simulation results of the Vendenheim demonstration site in MtCO₂
409 equivalents.**

410
411 The result of the Monte Carlo simulation is a mean value of 17.2 gCO₂e/kWh, and standard
412 deviation of 5.8 gCO₂e/kWh. The upper limit based on the 95th percentile is 19.3 gCO₂e/kWh,
413 and the lower limit based on the 5th percentile is 15.8 gCO₂e/kWh. These results were then
414 calculated with respect to the functional unit (1 kWh electricity produced) and combined with
415 the operational impact based on the two scenarios defined in the methods section. For scenario
416 1 (lower), the mean result is that the impact of the Vendenheim case study is expected to be
417 13.9 gCO₂e/kWh, with an upper limit based on the 95th percentile of 15.6 gCO₂e/kWh and a
418 lower limit based on the 5th percentile of 12.7 gCO₂e/kWh. The results for scenario 2 (upper)
419 reveal a considerably lower impact, with the mean expected to be 6.9 gCO₂e/kWh, with an
420 upper limit of 7.8 gCO₂e/kWh (95th percentile) and a lower limit (5th percentile) of 6.4
421 gCO₂e/kWh.

422
423 Figure 8 shows the distribution of impacts to different processes within the product system for
424 both scenarios as the difference between them is negligible. The result shows where most of
425 the impacts are sourced from, which can help identify opportunities for improvements.



426
427 **Figure 8. Estimated climate change impact of scenario 2 in the Vendenheim**
428 **demonstration site.**

429
430 As Figure 8 above shows, 50% of the climate change impact is derived from the steel material
431 used in the casing, followed by the mud products needed for drilling, then by the energy use in
432 drilling and the concrete. All these aspects are involved in the borehole construction phase of
433 the demonstration site. For the other impact categories, the same applies to particulate matter
434 formation and terrestrial acidification. However, the greatest human toxicity impacts are due to
435 the drilling waste and ionising radiation relating to the use of nuclear power plants to produce
436 the electricity used in drilling. However, when analysing the impact, it is interesting to see the
437 potential consequences of the co-product district heating in Vendenheim. If the district heating
438 supply is from the EGS power plant, it could substitute heating from other sources, e.g. natural
439 gas. It is assumed that the heating capacity factor is about 50%, as described in the methods
440 section. If natural gas utilisation is substituted, it could result in an avoided climate change
441 impact of about 304 gCO₂e for every 1 kWh of electricity that is produced, equal to about 15.9
442 Mt CO₂e on an annual basis.

443
444 To summarise the results of the life-cycle perspective for the Vendenheim case study, the most
445 critical impact categories are particulate matter formation, human toxicity, climate change,
446 ionising radiation and terrestrial acidification. As the study is conducted when the Vendenheim
447 demonstration site is still under construction, the impact analysis was conducted using Monte
448 Carlo simulations, and the uncertainty distribution was found using the Pedigree method, as
449 was conducted for the study of the Reykjanes demonstration site. The impact is then estimated
450 over a range for the two scenarios, giving results between 7.7 and 15.4 gCO₂e. Today, it is

451 considered more likely that the impact when operational will be closer to 11.0 gCO₂e based on
452 the currently expected flow rate and net electrical efficiency.

453

454

455 4. Discussion

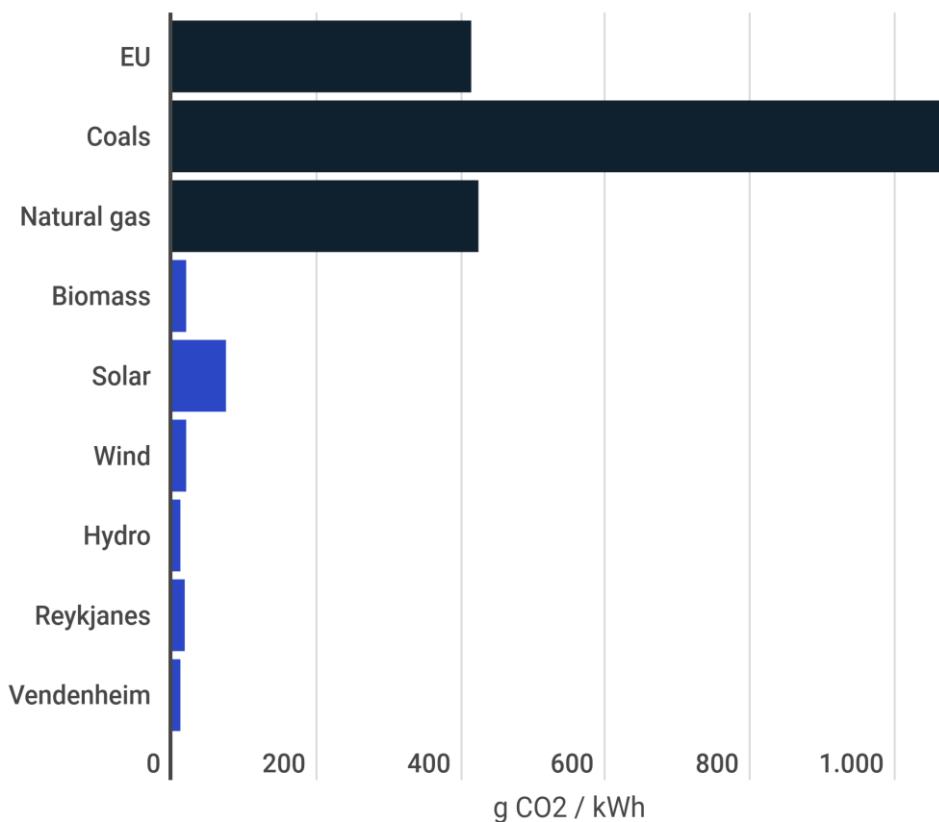
456

457 4.1 Comparison to climate impacts of alternative electricity-generating technologies and EU-
458 wide decision-making implications

459

460 The life-cycle perspective of the two demonstration sites was compared with the life-cycle
461 perspectives of various electricity production alternatives and the EU-27 average (EEA, 2020).
462 As shown in Fig. 9, the two demonstration sites have a fractional climate change impact from
463 a life-cycle perspective compared to non-renewable alternatives and a comparable effect to
464 renewable options such as hydropower, wind energy and biomass.

465



466

467

468

469 **Figure 9. Climate change impact of various electricity production alternatives and the
470 EU-27 average.⁴**

471

472

473

474 These results underscore the importance of electricity generation from EGS as a climate-
475 mitigating activity. It is even more important when considering that many of the renewable
476 alternatives are intermittent sources that can be hard to control without a storage system. EGS
477 electricity generation can support biomass utilisation and hydropower to provide renewable

⁴ Sourced from EEA (2020).

474 baseload in the EU's future renewable energy system. There is also the added benefit of being
475 able to provide heating as well as minimising reliance on natural gas or biomass.
476

477 In June 2020, a political agreement between the European Parliament Taxonomy Regulation
478 set in motion the imminent creation of a “green list” classification system for sustainable
479 economic activities. It details a general framework that will allow for the progressive
480 development of an EU-wide classification system for environmentally sustainable economic
481 activities (European Commission, 2020). The formalities of this general framework are,
482 however, still being developed and will take a few years. The proposed green list is as follows:
483 climate change mitigation, climate change adaptation, sustainable use and protection of water
484 and marine resources, a transition to the circular economy, pollution prevention and control,
485 and protection and restoration of biodiversity and ecosystems. Based on the low LCA impacts
486 of the demonstration sites at Reykjanes and Vendenheim, deep EGS ventures could provide a
487 significant contribution to at least two of these six objectives, i.e. climate change mitigation,
488 and pollution prevention and control, as well as fulfilling wider policy objectives linked to the
489 decarbonisation of economies, for example, the emissions reduction objectives of the Paris
490 Accord.

491
492 4.2 Further reducing the impacts of deep EGS power projects
493

494 Since the main impacts at the Reykjanes demonstration site, and other hydrothermal geothermal
495 energy sites, are greenhouse gas emissions from the borehole itself, these effects could be
496 largely avoided by carbon capture and mineralisation/utilisation, while the same approach is
497 also possible for hydrogen sulphide emissions. There are commercial solutions available that
498 have been tested for hydrothermal geothermal energy systems e.g. the CarbFix and Sulfix
499 projects at Hellisheiði in Iceland, where the gases react with basaltic sub-surface rocks to form
500 stable minerals for safe, long-term storage of the injected gases (Gíslason et al., 2010; Matter
501 et al., 2011; Sigfússon et al., 2018) and the Carbon Recycling International methanol plant
502 nearby at Svartsengi in Iceland, where renewable methanol from carbon dioxide and hydrogen
503 is produced using water electrolysis (Helgason et al., 2020; Stefánsson and Sigurbjörnsson,
504 2019).

505 The latter solution is using captured carbon dioxide emissions from another power plant
506 operated by HS Orka, and currently the CarbFix and Sulfix projects capture about 10,000 tonnes
507 of CO₂ and 5,000 tonnes of H₂S, respectively, injecting them back into the geothermal reservoir,
508 corresponding to 34% and 68% of the annual emissions from the plant in 2017 (Sigfússon et
509 al., 2018). The carbon available at these power plants differs as the amount of carbon emitted
510 from geothermal boreholes is very site-specific, but in general, the annual quantity is quite low
511 compared to natural gas or coal-fuelled power plants. This is a strength, since it is easier to store
512 small amounts than substantial quantities of carbon. However, carbon capture from
513 hydrothermal geothermal energy systems could also be combined with other utilisation
514 solutions that are also available e.g. bioenergy systems acting as a carbon boost to increase the
515 quality of the syngas or carbon, which could be used to make utilisation with captured carbon
516 more economically attractive than it is already by supplementing each other. Together,
517 bioenergy and geothermal energy could be part of the 100% renewable energy solution
518 (Connolly et al., 2016) in the future smart energy system (Sigurjónsson and Clausen, 2018).
519

520
521 In comparison to previous studies on EGS systems, the results at Vendenheim show lower
522 climate change impacts from a life-cycle perspective. The reason is that the borehole drilling
523 uses electricity as its primary driver. The power used in France has a relatively low carbon

intensity compared to the EU average (Ang and Su, 2016). Although it is mostly not from renewable sources, it has a significant impact on the overall impact of the Vendenheim demonstration site. The improvements available at Vendenheim and likely other hot, dry rock geothermal energy sites are mostly related to a transition to circular economy practices, as most of the impacts are from materials or energy used during construction. As noted above, sourcing the electricity used during drilling from low carbon sources is essential, and it is even better if it is sourced from renewable energy sources. This could be extended to adopt a circular supply chain, i.e. use of renewable energy, bio-based or potentially wholly recyclable materials. Reducing the impacts of the steel and cement used in the borehole construction is likely a challenge in the short to medium term. As the findings by Arens et al. (2017) suggest, efforts to reduce greenhouse gas emissions in the steel industry through new processes are unlikely to be available in time to help meet the United Nations' climate targets for 2030, as these technologies will take decades to develop and introduce. However, low carbon cement alternatives are already available, as noted by Naqi and Jang (2019), and should be considered for use in the geothermal energy sector.

The lower carbon intensity of deep EGS projects in comparison to fossil fuel alternatives and most other renewable energy solutions could increase the attractiveness of these innovative technologies as a means of decarbonising the French and Icelandic energy systems. In the case of the latter, the Icelandic electricity system is almost entirely fuelled by renewable energy, however, Iceland's new energy and climate policies stress the importance of further harnessing of renewable energy resources to provide electricity for the expanding electric car market and its charging infrastructure, and in ports (MENR, 2020; MII, 2020). Meanwhile, the results in this study provide further evidence in support of the contentions of Chavot et al. (2018) who asserted that deep EGS could form an important part of the energy transition in France, in which the carbon intensity of the electricity supply is further reduced in the period to 2030. This would help France to comply with its targets under the Paris Accord, which are to reduce non-Emissions Trading Scheme greenhouse gas emissions by 36% over the period 2005 to 2030 (European Commission, n.d.).

4.3 Study limitations

The main limitation of this study is that both case studies represented demonstration sites in the middle of construction. This means that actual data could not be used for parts of the assessment, requiring Monte Carlo simulation for the construction phase. However, this was mitigated by using estimates based on previous studies in a similar type of plant and area as the demonstration sites. The widely used Pedigree Matrix was used to estimate uncertainty for the input values that were assumed.

Additionally, several of the environmental and socio-cultural impacts of geothermal power projects were not captured in this assessment, most of which are traditionally overlooked in LCA studies. These include impacts to the aesthetics of geothermal areas, increased risk of land subsidence and earthquakes – as have recently been evident at the Vendenheim site (Think Geoenergy, 2020), noise emissions, and short and long-term biodiversity implications. These also need to be accounted for in decision-making, and this can be done using tools such as social-LCA, cost-benefit analysis, Environmental Impact Assessment and Multi-Criteria Decision Analysis (Cook et al., 2017; Cook et al., 2019).

574 **5. Conclusion**

575
576 The geothermal industry continues to grow in significance in many nations around the world.
577 Increasingly, EGS are utilised as a means of providing low carbon sources of energy,
578 contributing to the fulfilment of climate-related goals, such as those of the Paris Accord and the
579 Sustainable Development Goals, especially Goals 7 (Affordable and Clean Energy) and 12
580 (Responsible Production and Consumption). This study set out to provide the first LCA studies
581 in the academic literature pertaining to deep EGS, using the case studies of the DEEPEGS
582 demonstrator sites at Reykjanes, Iceland and Vendenheim, France to explore the climate change
583 impact of deep EGS projects, deciphering where in the life-cycle the impacts occur, and
584 reflecting on opportunities to mitigate greenhouse gas emissions.

585
586 The study found evidence that the greenhouse gas emissions of deep EGS projects across their
587 life-cycle are considerably lower than fossil fuel alternatives for electricity generation, and
588 comparable in emissions per unit of energy to renewable energy solutions such as hydropower,
589 wind energy and biomass. Climate change impacts for the sites were estimated to be in the
590 range 1.6-17.4 gCO₂e/kWh and 6.9-13.9 gCO₂e/kWh for Reykjanes and Vendenheim,
591 respectively. The main options for reducing greenhouse emissions at Reykjanes related to the
592 carbon capture and mineralisation of effluent carbon dioxide, as has been practiced elsewhere
593 in Iceland in recent years. The dry, hot rock project of Vendenheim had opportunities to reduce
594 greenhouse gas emissions by advancing its adoption of circular economy principles relating to
595 the procurement of materials, as most of the climate impacts derived from materials and energy
596 used in construction.

597
598
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