

1 **An environmental life cycle cost assessment of the costs of deep enhanced**
2 **geothermal systems – the case studies of Reykjanes, Iceland and**
3 **Vendenheim, France**

4
5
6 **Abstract**

7
8 Environmental life cycle costing (ELCC) is a tool which aggregates five categories of monetary
9 costs across a project's life cycle: investment, operation, maintenance, end-of-life, and
10 externalities. This paper summarises the results from the first two ELCC studies involving deep
11 enhanced geothermal systems (EGS). The ReCiPe method was used to transform life cycle
12 impact factors into economic costs for externalities. The two case studies are the pilot EGS
13 project in Reykjanes, Iceland and the Vendenheim co-generation plant in France. The ELCC of
14 the Reykjanes project is estimated in the range 14.47-15.78 million euros, with investment and
15 well drilling projected to constitute 83% of these amounts. An ELCC in the range 91.90-113.97
16 million euros is estimated for Vendenheim, with the production plant, well drilling, and
17 operations and maintenance costs accounting for the majority. The levelized costs of energy
18 associated with Vendenheim (mean €45.0/MWh/year) and Reykjanes (mean €16.5/MWh/year)
19 are at the lower end of the range normally reported for geothermal power projects. Although
20 the case studies cannot be directly compared since Reykjanes involves the drilling of a single
21 well and Vendenheim a co-generation plant and two wells, the outcomes suggest that deep EGS
22 projects may involve cost-savings compared to conventional geothermal power ventures.

23
24 **Keywords:** decision-making; valuation; deep geothermal energy; environmental impacts;
25 costing

1. Introduction

1.1 Life cycle costing and environmental life cycle costing

As a concept, life cycle costing (LCC) began in the mid-1960s when the US Department of Defence began an assessment of the long-term cost effects of military products (Lindholm et al., 2018). In recent years, the use of LCC has expanded in breadth considerably, especially among public authorities, who are often tasked with making cost-efficient investment decisions (Stark, 2015; Cheung et al., 2015; Kerzner and Kerzner, 2017; Kambanou et al., 2020). When comparing between options delivering the same project outcome, decision-makers are likely to prefer the least-cost or most cost-effective option (Stark, 2015). LCC seeks to provide a solid evidence base for making such decisions, endeavouring to account for all of the costs incurred by a project over its entire lifecycle (Jackson and Ostrom, 1980; Lindholm et al., 2017; Heralova, 2017). In general, conventional LCC methods account for and aggregate four categories of costs: investment, operation, maintenance and end-of-life disposal (Luttenberger and Luttenberger, 2016; Heralova, 2017). LCC has been defined by the International Organization for Standardization standard, Building and Constructed Assets, Service-life Planning, Part 5: Life-cycle Costing (ISO 15686-5) as an “*economic assessment considering all agreed projected significant and relevant cost flows over a period of analysis expressed in monetary value. The projected costs are those needed to achieve defined levels of performance, including reliability, safety and availability.*” A broader perspective than LCC is offered through environmental life cycle costing (ELCC). This method accounts for all of the four categories of costs addressed by LCC, but also the external environmental costs of a project across its lifecycle (Swarr et al., 2011).

Accounting for such impacts has become an increasingly critical component of decision-making in relation to the award of public procurement contracts in the EU. Article 67 of the EU Procurement Directives¹ asserts that contracting authorities must base the award of public contracts on the most economically advantageous tender, using a cost-effectiveness approach such as life cycle costing (Luttenberger and Luttenberger, 2016). The best ‘price-quality ratio’ of various tenders may be applied as a decision-making criterion, to include the costs of their environmental and social implications. Consequently, Article 67 stimulates a subtle but discernible shift in decision-making emphasis – from evaluating project tenders based on ‘lowest price’ to ‘lowest cost’ (Van den Abeele, 2014). In addition, the EU Procurement Directives have further embedded the importance of environmental soundness into decision-making processes by stipulating that project economics cannot be the only criterion contracting authorities apply. The award of contracts can also be based on various criteria of relevance to life cycle analysis (LCA), including production processes, sourcing of raw materials, water or energy consumption, or biodegradability (ICLEI, n.d.). Article 68 of the EU Procurement Directives further reinforces the importance of the LCC concept in public procurement, specifically enshrining the need to account for environmental externalities across a project’s lifecycle, such as pollution, greenhouse gas emissions and recycling costs, provided their monetary value can be determined and verified.

1.2 LCA studies on geothermal power generation

For well over 100 years, geothermal resources have been harnessed to provide base-load electricity around the world (Tomasini-Montegro et al., 2017). Globally, the total installed

¹ 2014/23/EU (the Concessions Directive), 2014/24/EU (the Public Sector Directive) and 2014/25/EU (the Utilities Sector Directive)—hereafter simply referred to as the 2014 Directives.

95 capacity from geothermal technologies increased by 3.649 Gigawatt equivalent (GWe) in the
96 period 2015-2020, a 27% increase. A further increase of 19% is projected over the period 2020-
97 2025 (Huttrer, 2020). Given the likely expansion in production, increasingly attention has been
98 directed towards the environmental impacts of the geothermal power sector, particularly in
99 connection with the harnessing of high-temperature fields for the purposes of electricity
100 generation. In addition, there are also environmental impacts from exploration and some natural
101 background emissions of greenhouse gases and hydrogen sulphide, from geothermal areas
102 (Ármansson, 2018; Bayer et al., 2013; Paulilo et al., 2019; Santoyo et al., 2018). A variety of
103 studies have been undertaken to explore the environmental impacts of geothermal power
104 projects and their impacts on societal well-being (Bayer et al. 2013; Shortall et al., 2015; Cook
105 et al., 2020). Bayer et al. (2013) articulate that these include direct effects, such as land
106 distortion, production of geological hazards, atmospheric emissions, waste heat, solid waste,
107 water consumption, noise emissions, and impacts on biodiversity. However, in addition to these
108 impacts, there exist a variety of indirect effects specific to the harnessing of geothermal power.
109 These occur in relation to the materials and energy required over the lifecycle of the power
110 plant (Tomasini-Montegro et al., 2017). Therefore, in order to fully understand the
111 environmental implications of geothermal power generation, a product system approach should
112 be favoured, since this facilitates an assessment that is considerate of the supply chains of the
113 life cycle in this industrial sector. In so doing, information is provided about the environmental
114 impacts pertaining to the whole system, as opposed to a single process in the geothermal
115 industry. Evaluations of this type can be conducted using LCA, which, when conducted
116 according to well-defined ISO norms and processes (such as ISO 14040, 2006a; ISO 14044,
117 2006b), can compile all inputs, outputs and environmental impacts of geothermal power
118 generation across a project's lifecycle (Hunkeler et al, 2008; Swarr et al., 2011).

119
120 In recent years, with greater recognition of the need to tackle climate change, the many direct
121 and indirect environmental impacts of power generation have assumed increased importance.
122 Even fairly early examples in the academic literature hint at a breadth of focus in LCA studies
123 with regards to the power sector. They include a study on the greenhouse gas emissions of
124 substituting switch grass for coal in electricity generation (Ney and Schnoor, 2002), an LCA on
125 a natural gas combined-cycle power system (Spath and Mann, 2000), and an LCA of a wind
126 turbine (Batumbya et al., 2006). Several LCA studies can be found in the academic literature
127 that have focused on geothermal power plants. These include but are not limited to:

- 128
- 129 • A theoretical study on the potential LCA impacts of EGS (Clark et al., 2012);
- 130 • An applied LCA study on geothermal power generation using supercritical steam (Frank,
131 et al., 2012);
- 132 • A review study on the environmental issues of relevance to geothermal LCAs in the
133 context of a Californian plant (Sullivan et al., 2012);
- 134 • An analysis of how to integrate life cycle analysis and energy synthesis in respect of a
135 dry steam geothermal plant in Italy (Buonocore et al., 2015);
- 136 • A comprehensive review of existing LCA studies using different geothermal
137 technologies (Tomasini-Montenegro et al., 2017);
- 138 • A study focused on climate change impacts in relation to a power project in the Upper
139 Rhine Valley (Pratiwi et al., 2018);
- 140 • An LCA on a geothermal power plant linked to the Southern German Molasse Basin's
141 hydrothermal resource (Menberg et al., 2021).
- 142 • An LCA assessment on the combined heat and power double-flash geothermal power
143 plant of Hellisheiði in Iceland (Colucci et al., 2021).

- The LCA studies on the deep enhanced geothermal systems of Reykjanes and Vendenheim, on which these LCC studies are largely based, are reported in Sigurjónsson et al. (2021).

1.3 LCC and ELCC studies on geothermal power generation

In general, the economic efficiency of power plants continues to be assessed using standard cost-benefit analysis, which is not extended to account for the economic costs of environmental and social effects of projects (Cook et al., 2016), and levelized cost analysis, which often suffers from the same pitfall (Ebenhoch et al., 2015). It is likely that this is largely due to reasons of complexity deriving from the challenges in (a) estimating the costs of projects with a long lifespan, and (b) converting environmental impact factors into economic cost values. Very few LCC or ELCC studies, which seek to incorporate the economic value of environmental externalities, have occurred linked to the harnessing of geothermal energy resources. The paper by Martínez-Corona et al., (2017) appears to provide the first study which has sought to link LCA inventories and impact values to monetary valuation in the context of high-temperature geothermal power plants. Based on a power project in Wairakei, a comparison was formed concerning different environmental assessment methods for geothermal plants based on either physical or monetary data. Yilmaz (2020) conducted both an LCC and levelized cost assessment of hydrogen production via geothermal power. Park et al., 2020 explore the impact of risk on the levelized cost of geothermal technologies. Other life cycle cost studies in the context of geothermal power have focused on low temperature geothermal technologies, such as ground source heat pumps (Chiasson, 2006; Habibzadeh-Bigdarvish et al., 2019).

1.4 Deep EGS and the DEEPEGS case studies of Reykjanes, Iceland and Vendenheim, France

In theory, drilling to unconventional depths in EGS projects enables the sourcing of water at 400-500°C, some 25-50% hotter than in conventional hydrothermal drilling projects in volcanic regions (Friðleifsson et al. 2018; Friðleifsson et al., 2020). The higher temperatures potentially associated with EGS projects could be used to generate superheated steam, which can then be utilised to create electricity. Assuming success, an anticipated consequence of deep drilling projects is a considerable reduction in costs and environmental impacts, since far fewer wells would be required to be drilled to generate the same power output using regular geothermal boreholes. Since the harnessing of geothermal energy can result in changes to land surface manifestations, noise effects and visual blight from pipelines (Shortall et al., 2015; Cook et al., 2020), the drilling of fewer wells, including make-up wells, would reduce the scale of these effects and the environmental impacts of drilling. In this respect, this ELCC study acts as a starting point in providing a comparison in cost per unit of power output between conventional and EGS projects.

DEEPEGS is an innovative four-year project led by the Icelandic power company, HS Orka, alongside other partners in Iceland, France, Germany, Italy and Norway. The project includes the testing of deep EGS projects in Reykjanes, Iceland and Vendenheim, France. An existing 2.5 km deep well at Reykjanes, known as IDDP-2, has been deepened to 4.6 km, with drilling completed in late 2017, and well stimulation and testing undertaken thereafter. The Vendenheim dry steam project is led by Fonroche Géothermie and involves the drilling of two deep wells in excess of 5,000 metres depth, with side-tracks starting below a depth of 3,500 metres to increase connectivity between the VDH1 and WDH2 wells, and the subsequent construction of power plant infrastructure on the site of a former fossil fuel plant. Unlike Reykjanes, where the aim was to provide additional electricity generation, the Vendenheim

194 project entailed the co-production of heat and power from a geological area with heat in the
195 range of 180-210 °C from the deep wells (Sanjuan et al., 2020). Although, these are much lower
196 temperatures than at commensurate depths in Iceland, in the context of mainland Europe, where
197 there is different geology, the Vendenheim site is a very hot resource. Thus, site-specific
198 differences make it difficult to compare the two case studies.

199

200 1.5 Aims, research questions and structure

201

202 The limited number of LCA/LCC/ELCC studies in the academic literature with respect to
203 geothermal power projects, and complete absence in the context of deep EGS ventures,
204 motivates this study. There are thus considerable opportunities to enhance knowledge
205 concerning the lifecycle costs of different types of geothermal power projects. This will enable
206 better knowledge of the lifecycle costs of geothermal power compared to alternative energy-
207 generating technologies and ultimately help to inform cost comparisons between different
208 geothermal technologies and design parameters. In particular, the highlighting of the economic
209 value of environmental hotspots across the lifecycle of geothermal power project provides
210 important information for supply chain and project managers.

211

212 The aim in this study is not to form direct cost comparisons between the two case studies, since
213 one is a new project and the other an extension of an existing venture, but rather to add to the
214 knowledge base concerning the economic costs of geothermal power ventures. It applies the
215 two cases studies in pursuit of answers to the following three research questions:

216

- 217 1) What are the environmental life cycle costs and how do these affect the levelized costs
218 of deep EGS power projects?
- 219 2) What represent the main cost components in deep EGS power projects?
- 220 3) How do the environmental life cycle costs of deep EGS power projects compare to
221 conventional geothermal power ventures?
- 222 4) What other environmental and socio-cultural impacts to human well-being should be
223 accounted for beyond those incorporated in the ELCC?

224

225 This paper applies the LCA model (Sigurjónsson et al., 2021) developed in relation to the
226 DEEPEGS site at Reykjanes and Vendenheim to conduct an ELCC, translating environmental
227 impacts into monetary values using state-of-the-art methods. Section 2 of this paper sets out the
228 methodology for the ELCC study on the DEEPEGS project in Reykjanes, Iceland, including
229 how it accounts for environmental externalities and conducts a sensitivity analysis concerning
230 the discounting of probabilistic cost components. Section 3 sets out the estimated results
231 according to the range of discount rates explored, and then discusses the main outcomes from
232 the study, including some of the wider decision-making implications of LCC, and the extent to
233 which the approach accounts for all environmental and social costs linked to the project. Section
234 4 provides a brief conclusion and considers the potential for future related research.

235

236 2. Methodology

237

238 2.1 Components of ELCC

239

240 Within LCC and ELCC, some costs can be classed as deterministic and others probabilistic.
241 Deterministic costs include mainly those outlays relating to acquisition and disposal. Most
242 probabilistic costs relate to the reliability and maintenance needs of a project's system,
243 including the costs of repairs, spare parts and downtime (Barringer, 2003). This paper adopts

244 the following seven-component formula for calculating the ELCC of the DEEPEGS case
245 studies at Reykjanes and Vendenheim, as set out in Fig. 1. Unlike traditional LCC, ELCC
246 incorporates an additional category of costs involving environmental externalities linked to the
247 operations of a project, for instance greenhouse gases and other pollutants (Luttenberger and
248 Luttenberger, 2017). ELCC thus accounts for both the internal and external environmental
249 externalities of a project. It is calculated as follows:

$$250 \\ 251 \text{ELCC} = C_{pio} + C_{wd} + C_{pp} + C_{ret} + C_{de} + C_{om} + C_{ex} \quad (\text{eq. 1})$$

252
253 Where:

254
255 ELCC = the aggregate environmental life cycle costs expressed in present value and across the
256 likely lifespan of the project (30 years)

257 C_{pio} = Project planning, upfront infrastructure and operational

258 C_{wd} = Well drilling

259 C_{pp} = Production plant

260 C_{ret} = Reservoir engineering (including stimulation) and testing

261 C_{de} = Decommissioning

262 C_{om} = Operations and maintenance

263 C_{ex} = Environmental externalities of operations

264

265 Note that not all the cost elements in equation (1) applied to both case studies. For example,
266 there was no production plant (C_{pp}) in relation to the Reykjanes site, as one with 100 MW_e
267 capacity already operated at the site, drawing energy from existing multiple conventional wells
268 of 2.2-2.5 km depth. The DEEPEGS well at Reykjanes was the first deep well drilled into the
269 deeper, higher-temperature zone.

270

271 Physical impact data was extracted from LCA studies on the two case studies conducted by the
272 University of Iceland (Sigurjónsson et al., 2021). The system boundaries and constraints for the
273 ELCC studies are set in accordance with the LCA studies reported by Sigurjónsson et al. (2021).
274 The system boundaries related to costs and impacts occurring at the construction site, except
275 for externalities, which constitute spill over effects on surrounding populations. Aggregate cost
276 data for the various deterministic and probabilistic cost components (excluding environmental
277 externalities) was provided by the developers responsible for the respective sites. The exception
278 to this was the cost in respect of decommissioning in the case of the Reykjanes project, where,
279 in the absence of a developer estimate, ‘rule of thumb’ data was applied from Huenges et al.
280 (2010).

281

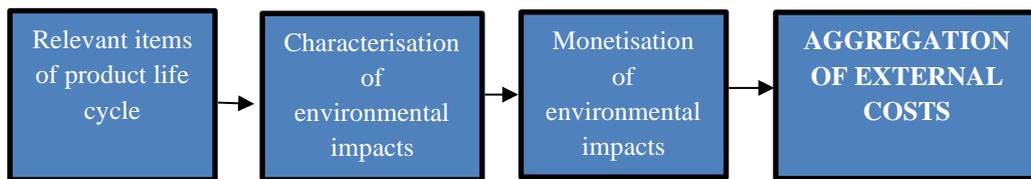
282 2.2 Accounting for environmental externalities of operations

283

284 The concept of external costs, or negative externalities, was first introduced by the British
285 economist, Arthur Pigou, who sought to address such effects through market mechanisms,
286 including environmental taxes. Rebitzer and Hunkeler (2003, p.253) define externalities as
287 costs that *include “the monetised effects of environmental and social impacts not directly billed*
288 *to the firm, consumer, or government, etc. that is producing, using, or handling the product.*
289 *The so named “externalities” are outside the economic system, though inside the natural and*
290 *social system”. In a lifecycle costing context, environmental externalities include costs already*
291 *paid for by an entity along the value chain and not included in market transactions (e.g.*
292 *municipal waste disposal, increased safety features), as well as costs that have been paid for*
293 *and can be monetised (e.g. greenhouse gas emissions and pollutants) (Hunkeler et al., 2008).*

294 As the results and discussion section of this paper acknowledges, there are also external costs
295 that are difficult to monetise using lifecycle costing, such as impacts to the aesthetics of
296 geothermal landscapes.

297
298 Several different techniques can be applied to account for environmental externalities. One
299 approach to calculating environmental externalities specific to life cycle costing is set out in the
300 European Union’s Clean Vehicles Directive (2009/33/EC). This Directive requires contracting
301 authorities to account for energy consumption, greenhouse gas emissions and pollutant
302 emissions. A set of standardised costs for carbon dioxide, nitrogen oxide, non-methane
303 hydrocarbon and particulate matter emissions from road transport is listed in the Appendix to
304 the Directive, which requires adjustment to reflect current rather than 2007 prices. More
305 recently, the European Commission has issued a new LCC Tool known as ReCiPe, inclusive of
306 direct and some indirect (environmental externalities) costs. ReCiPe first translates the relevant
307 items of a product’s lifecycle (e.g. its electricity consumption) into a resource and emissions
308 profile using publicly available life cycle inventory data, converts these environmental impacts
309 using LCA, before finally applying monetisation factors, such as the social cost of carbon and
310 damage function, to the computed environmental impacts and aggregating (Estevan and
311 Schaefer, 2017) (see Fig. 1). External costs were calculated using monetary impact values from
312 several air pollutants (Gunnlaugsson et al., 2018; Sigfússon et al., 2018) and a social cost of
313 carbon study (Sigfússon et al., 2018) focused on geothermal power, which assumes a value of
314 20 euros/tonne.



315
316
317
318
319
320
321
322 **Fig. 1. EU’s ReCiPe method to account for environmental externalities.**

323 324 325 2.3 Discounting for probabilistic costs and sensitivity analysis

326
327 An important issue in any ELCC study concerns the treatment of probabilistic costs relating to
328 costs not occurring upfront. The academic literature includes several studies focused on the
329 optimisation of probabilistic components, with much of the focus on operations and
330 maintenance costs (Tsang et al., 2006; Campbell et al., 2011). However, in an ELCC context,
331 much less attention has been given to the importance of discounting probabilistic costs, most
332 likely due to the issue of uncertainty. In order to be able to aggregate and compare costs that
333 occur at different points in the lifecycle of projects, these need to be made time-equivalent. This
334 is even more important when projects have an extended lifecycle, such as infrastructure
335 ventures. Discounting accounts for the time value of money (Jaggi et al., 2016; Chakrabarty &
336 Chaudhuri 2017). In LCC or ELCC studies, discounting converts future probabilistic costs into
337 present values (Goh and Sun, 2016; Islam et al., 2016). This enables future costs to be compared
338 with upfront outlays, and an overall aggregation of total costs to occur in present value terms.
339 The interest rate selected to discount future costs is an arbitrary value but is most often selected
340 based on an investor’s opportunity cost of money over time (Ciambrone, 2018). In other words,
341 the return that an investor wants to achieve is at least as high as the next best alternative
342 investment. In this paper, three discount rates are applied to the probabilistic costs in the

343 DEEPEGS ELCC model, including 2% (low), 5% (medium) and 8% (high). This sensitivity
 344 analysis ensures that the results account for various degrees of risk and uncertainty.

345
 346 In addition, a further sensitivity analysis was conducted with respect to the anticipated social
 347 cost of carbon in future years. The same approach was followed as per the study by Helgason
 348 et al. (2020). Although the starting point social cost of carbon was 20 euros per tonne as per
 349 Sigfússon et al. (2018), cost curves for future years then followed assumptions of low, medium
 350 and high per annum annual increments across the thirty-year period of this study. The
 351 increments are in accordance with estimates by the EIB (2015) and are set out in Table 1.
 352 Depending on whether low or high annual increments are assumed, the range of social carbon
 353 estimates in the year 2050 are from a low of 50 euros/tonne up to 200 euros/tonne.

355 **Table 1. Incremental increases in shadow prices for greenhouse gas emissions from 2020-2050**
 356 **(2018 euros) (sourced with permission from Helgason et al., 2020).**

		Time period		
		2018-2030	2031-2040	2041-2050
Low cost increase p.a.	Euros	0.5	1	2
Central cost increase p.a.	Euros	1	2	4
High cost increase p.a.	Euros	2	4	8

357

358

359 **3. Results and discussion**

360

361 **3.1 Results**

362

363 Table 2 summarises the ELCC calculation for the two case studies. A breakdown of the
 364 aggregate costs is detailed according to the various components provided in the methodology,
 365 including the respective percentage contributions. Figs. 2 and 3 display pie charts relating to
 366 these cost components. Both Table 2 and Figs. 2 and 3 are based on an assumed discount rate
 367 of 5% for all probabilistic cost components, including future costs relating to the imminent well
 368 stimulation, and these were assessed across a 30-year time horizon. Process externalities for the
 369 well drilling and stimulation components include all production and transport-related emissions,
 370 and these are derived directly from the LCA study on this project. Assessed environmental
 371 externalities for Reykjanes include carbon dioxide emissions (CO₂) and hydrogen sulphide
 372 emissions (H₂S). CO₂ and H₂S emissions, although much lower in scale than fossil-fuel
 373 alternatives, are considered to be by far the largest constituents of all steam vapour gases
 374 emitted by geothermal power plants, in Iceland and around the world (Júlíusson et al., 2015;
 375 Sigfússon et al., 2018). No fugitive emissions are assumed in relation to the operations of the
 376 Vendenheim project.

377

378

379

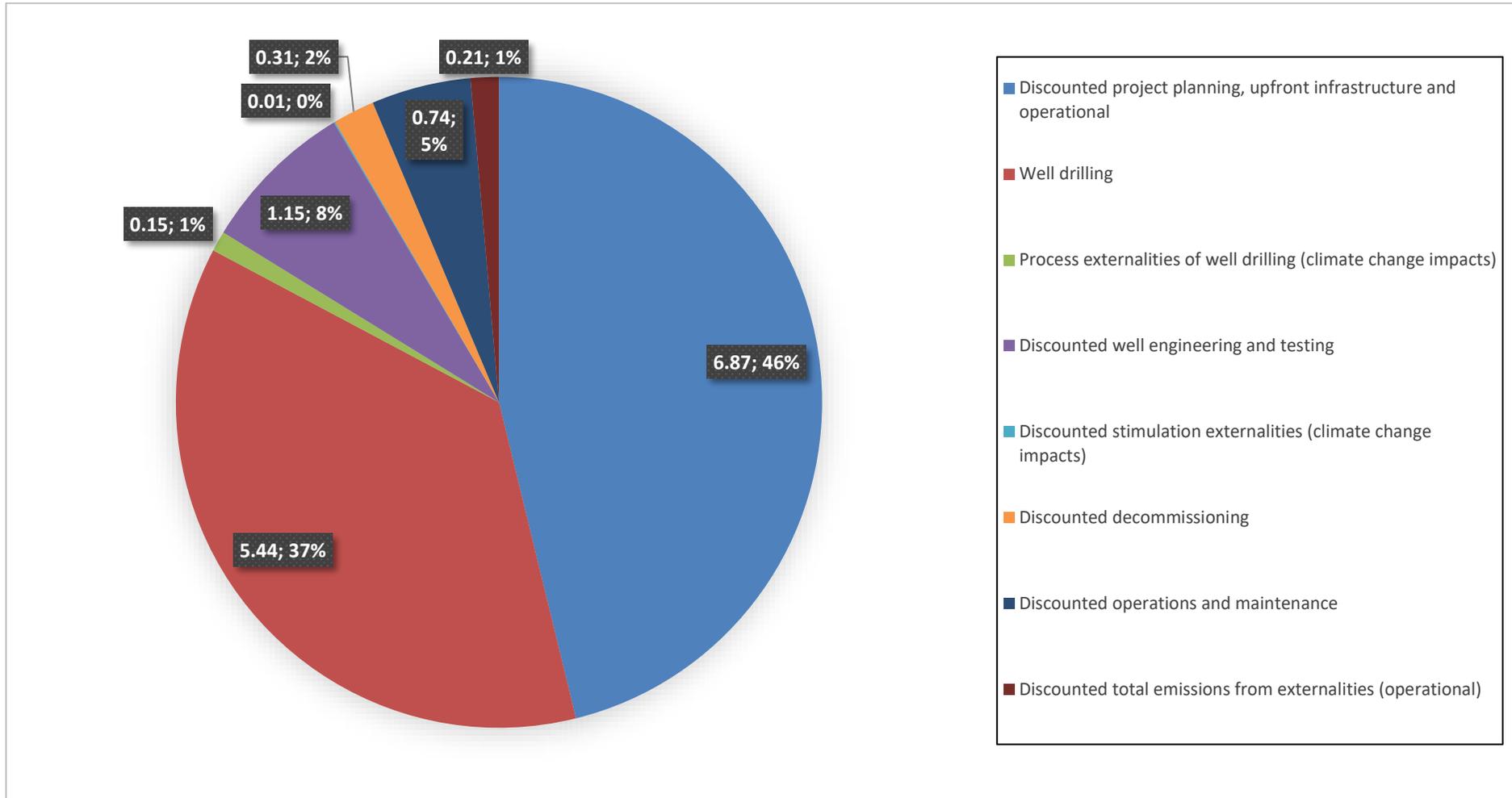
380

381

Table 2: ELCC calculation – Reykjanes and Vendenheim

Cost Component	Reykjanes		Vendenheim	
	Cost (million euros)	Percentage of total costs excluding environmental externalities of operations	Cost (million euros)	Percentage of total costs excluding environmental externalities of (operations
Project planning and upfront infrastructure and operational	6.87	46.83	5.40	5.39
Well drilling	5.44	37.09	31.52	31.43
Construction-related climate impacts (includes well drilling and stimulation	0.15	1.08	0.03	0.03
Production plant	0.00	0.00	34.30	34.20
Discounted well engineering and testing	1.15	7.83	2.57	2.57
Discounted decommissioning	0.31	2.13	0.50	0.50
Discounted operational and maintenance	0.74	5.02	25.97	25.89
Total costs excluding environmental externalities of operations	14.67	100.00	100.29	100.00
Environmental externalities of operations	0.21		0.0	
Total costs including environmental externalities of operations	14.88		100.29	

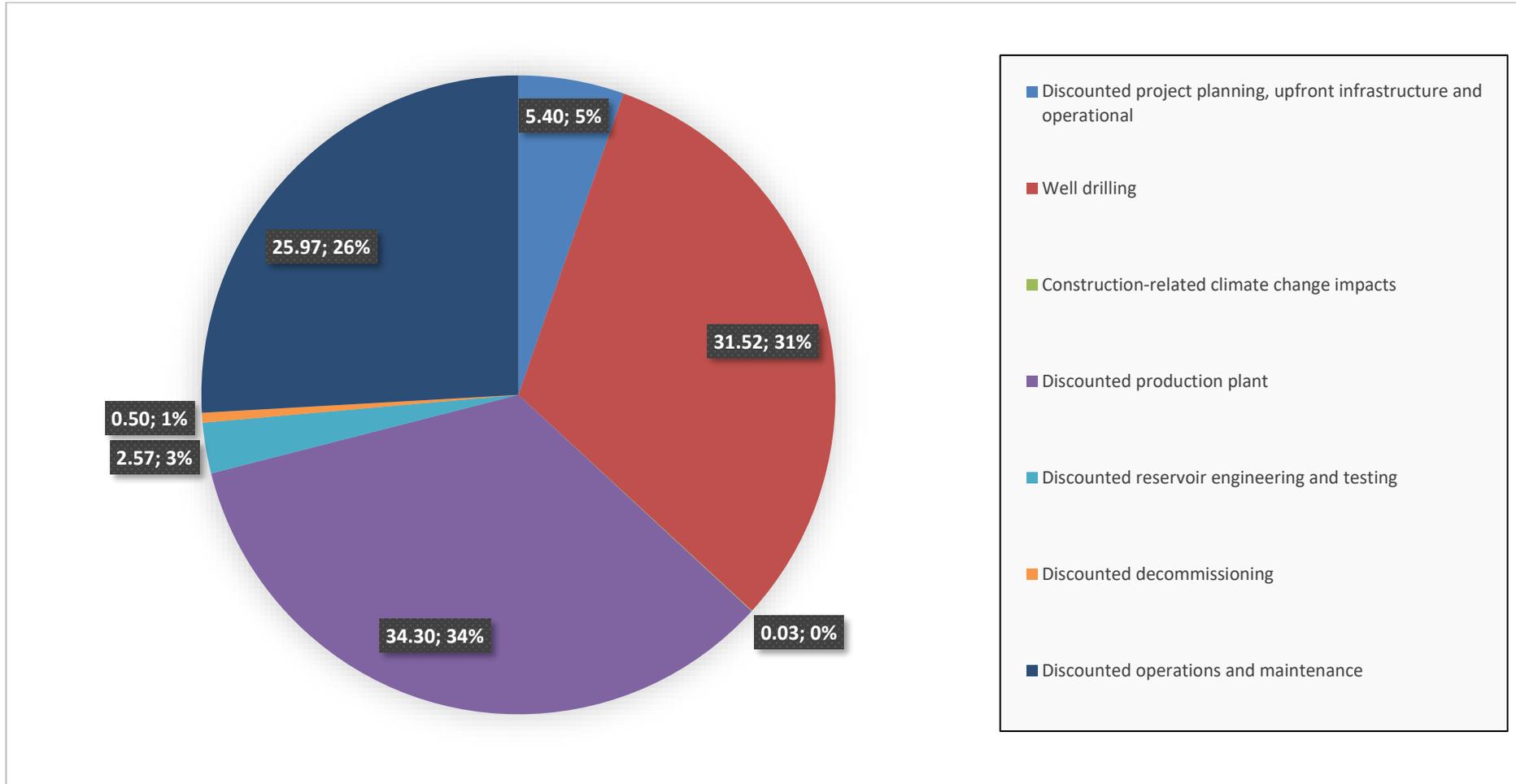
382



384
 385
 386
 387
 388
 389

Fig. 2: ELCC pie chart – breakdown of cost components, Reykjanes (million euros)

390
391
392



393
394

Fig. 3: ELCC pie chart – breakdown of cost components, Vendenheim (million euros)

395 The total life cycle costs (excluding environmental externalities from operations) for Reykjanes
 396 are 14.67 million euros. Including the environmental externalities of operations, this figure is
 397 marked up by 1.43% to 14.88 million euros. Given the capital intensity of all geothermal power
 398 projects and the relatively low greenhouse gas emissions in comparison to fossil fuel
 399 alternatives, it is perhaps unsurprising that environmental externalities of operations form a
 400 relatively insignificant component of the ELCC. Of the aggregate present value of 0.21 million
 401 euros (based on a 5% discount rate) ascribed to the environmental externalities of operations,
 402 these were dominated by greenhouse gas emissions associated with the well drilling phase (0.15
 403 million euros), especially relating to transportation of materials and equipment.

404
 405 A significant proportion (83%) of the total lifecycle costs at Reykjanes related to the well
 406 drilling phase and linked upfront costs concerning the procurement of equipment, labour, travel
 407 and miscellaneous components. This is a high proportion, but broadly in line with estimates in
 408 the academic literature given the specifics of this project and its speculative nature. The study
 409 by Huenges and Ledru (2010) found that 74% of the overall lifecycle costs of a geothermal
 410 power plant related to well drilling and investments linked to well drilling. Overall, probabilistic
 411 cost estimates constituted a slight contribution to the aggregate costs compared to deterministic
 412 components. The combined end of life, operations and maintenance, and environmental
 413 externalities of operations costs were 1.26 million euros, equating to 8.47% of the total costs
 414 inclusive of environmental externalities. Although the evaluation by Huenges and Ledru (2010)
 415 did not seek to account for the environmental externalities of operations, much higher
 416 operations and maintenance and decommissioning costs are associated with geothermal plants,
 417 compared to a single well, and accordingly, these reduce the proportion of total lifecycle costs
 418 derived from well drilling.

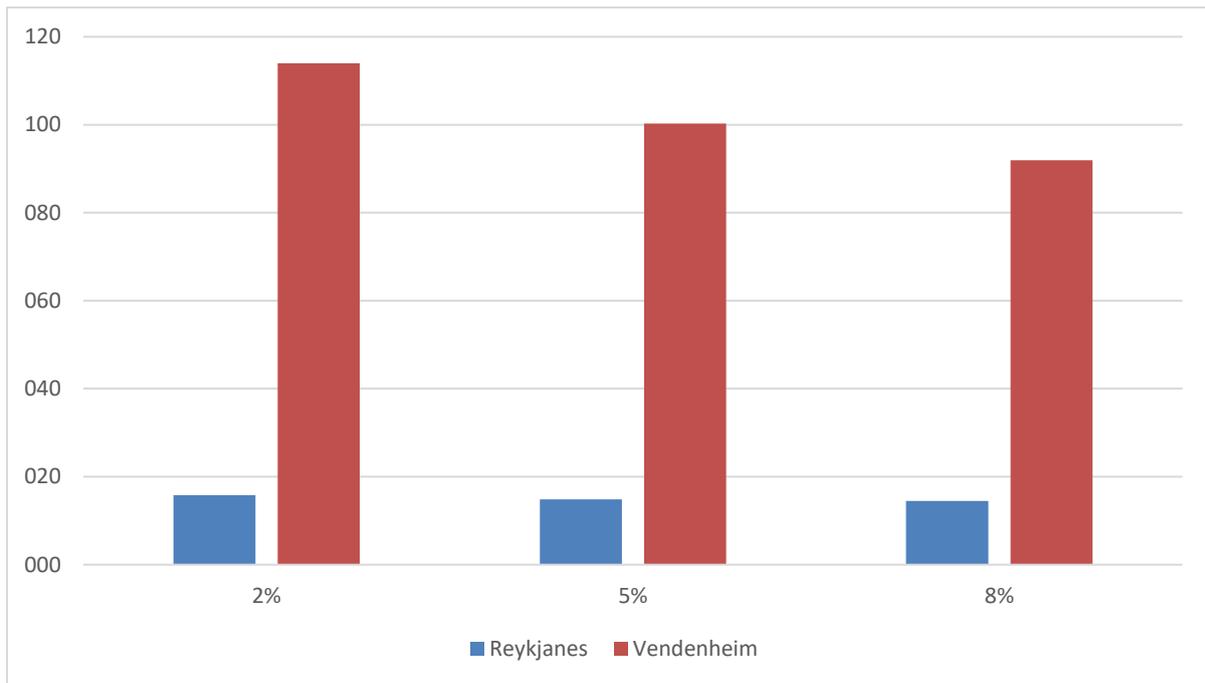
419
 420 The total life cycle costs for Vendenheim are 100.29 million euros. The results for Vendenheim
 421 are fairly typical of geothermal power projects, with 73% of costs amassed in relation to upfront
 422 activities, well drilling and the production plant. Construction-related externalities formed a
 423 very low proportion (0.03%) of the total estimated costs. Probabilistic cost estimates constituted
 424 26.47 million euros in relation to decommissioning and operations and maintenance. These two
 425 elements were equivalent to 26.39% of the total costs of the Vendenheim project, dominated
 426 by the 25.97 million euros of discounted costs in relation to operations and maintenance.

427
 428 Table 3 outlines the range of likely ELCC estimates according to the three selected discount
 429 rates of 2% (low), 5% (medium) and 8% (high). Fig. 4 provides a bar chart comparing these
 430 outcomes for both Reykjanes and Vendenheim.

431
 432 **Table 3: ELCC cost estimates (million euros) according to sensitivity analysis –**
 433 **DEEPEGS, Reykjanes**

Discount rate	Reykjanes		Vendenheim	
	ELCC (exc. operational environmental externalities)	ELCC (inc. operational environmental externalities)	ELCC (exc. operational environmental externalities)	ELCC (inc. operational environmental externalities)
2%	15.48	15.78	113.97	113.97
5%	14.68	14.88	100.29	100.29
8%	14.31	14.47	91.90	91.90

434
 435



436
437

Fig. 4: Bar chart comparing ELCC estimates (including environmental externalities) according to sensitivity analysis – Reykjanes and Vendenheim (million euros)

439

440

441

442

443

444

445

446

447

448

449

450

The relatively small contribution of probabilistic cost elements ensures that the choice of discount rate has a relatively insignificant effect on the outcomes in the sensitivity analysis. Based on the discount rates applied, the total ELCC for Reykjanes are in the range 14.47 to 15.78 million euros. There is a difference of 1.31 million euros (9.05%) between these estimates. The total ELCC for Vendenheim is in the range 91.90 to 113.97 million euros depending on the respective discount rate. This is a differential of 22.07 million euros (24.02%). The greater variance in the ELCC estimate for Vendenheim can be attributed mainly to the higher proportion of costs relating to probabilistic cost elements, especially the operations and maintenance of the new power plant facility.

451

452

453

454

455

456

457

458

459

460

461

462

463

464

465

Finally, a further sensitivity analysis was conducted by varying the social cost of carbon in accordance with the annual increments set out in Table 1. Based on the highest increments estimated by EIB (2020), the ELCC for Reykjanes, inclusive of operational externalities, increases to the range 15.13-17.05 million euros. The contribution of environmental externalities to the total was 1.57 million ISK in the scenario of lowest discount rate (2%) and highest social cost of carbon (average of 125 euros per tonne across the thirty-year horizon). This was equivalent to 9.21% of the total ELCC, an increase on the 1.94% contribution to the total ELCC (2% discount rate) when a non-varying social cost of carbon of 20 euros/tonne was assumed across the thirty-year horizon. There were no implications for Vendenheim’s ELCC due to the zero greenhouse gas emissions for the operations phase in the LCA by Sigurjónsson et al. (2021), and thus it is not possible to compare the relative effects of this part of the sensitivity analysis between the two study sites.

3.2 Comparison of outcomes to similar studies

466

467

468

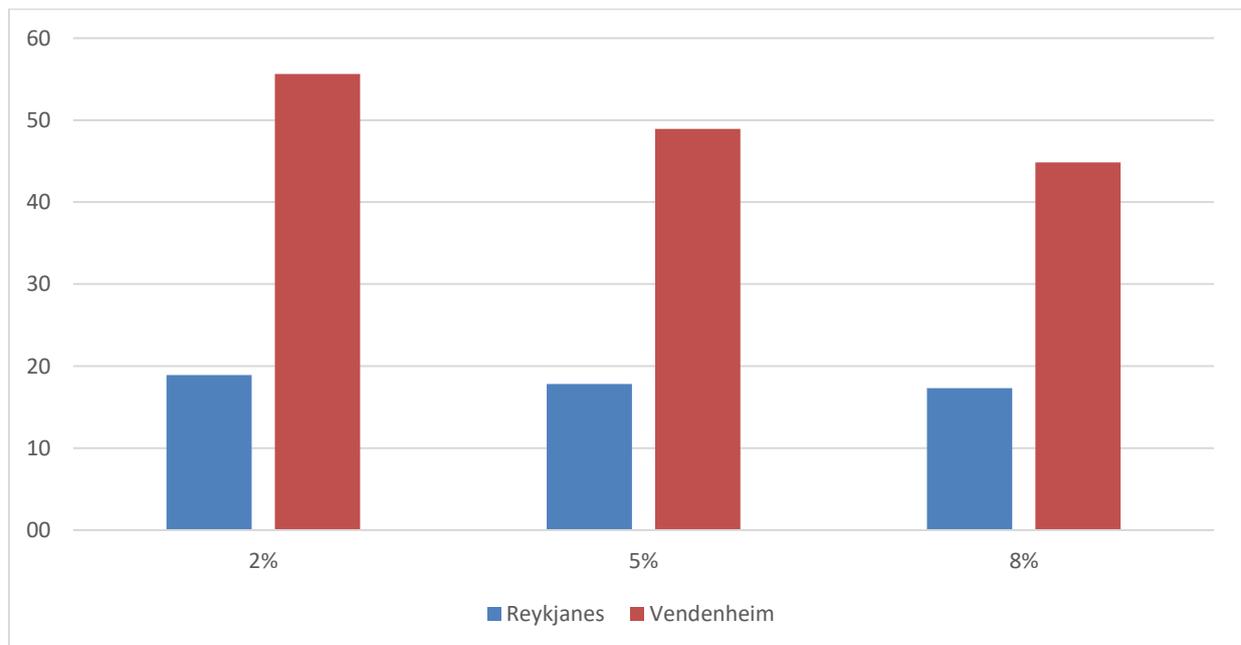
There are currently no comparable ELCC studies related to a single geothermal well, such as Reykjanes, in existence. Although other studies, such as Huenges and Ledru (2010), have provided indicative rule-of-thumb estimates in relation to geothermal power plants deriving

469 from EGS, those outcomes cannot be directly compared to this study as they include very large
470 costs related to power plant infrastructure, ongoing operations and maintenance costs, and they
471 also assume the drilling of multiple wells.

472
473 An additional way of considering the outcomes from this study is to consider these in terms of
474 the levelized cost of energy (LCOE), where multiple studies have focused on the economic
475 costs of geothermal power. LCOE is an economic measure describing the cost of generating a
476 MWh of electricity averaged over the lifetime of the power plant or energy-generating
477 technology. In other words, it averages the total ELCC over the production period, which in
478 this study was assumed to be 30 years.

479
480 The LCOE for the Reykjanes and Vendenheim projects were estimated to be €16.5/MWh/year
481 and €45.0/MWh/year, respectively, based on a 5% discount rate (see Fig. 6). Considerable
482 caution should be expressed with regards to these estimates, given that they were made prior to
483 commencement of production at both sites. Thus, various assumptions had to be made, such as
484 those relating to power outputs, capacity factors and future variable costs. Although it was
485 largely self-evident that the LCOE for the deep geothermal well at Reykjanes would be lower
486 than usual estimates for geothermal power projects, the same is also true – to a much lesser
487 degree – for the co-production project at Vendenheim. Recent estimates of the LCOE for
488 geothermal power, based on no subsidies, are in the range of US\$ 71-111/MWh/year, which
489 equates to approximately € 65-102/MWh/year (Lazard, 2018). Another study by IRENA (2018)
490 arrived at a LCOE of geothermal power estimate in the range 40 to 140 US \$ per MWh per
491 annum (approximately € 37-94/MWh/year, and thus, the Vendenheim estimate is at the lower
492 end of the reported range. The outcome also compares favourably with LCOE estimates for
493 geothermal power projects in Iceland, which have been estimated in the range € 37-
494 92/MWh/year (Ólafsson, 2016). The estimate for Reykjanes is even lower, but it is
495 incomparable to other estimates in the literature, since it did not include the costs of production
496 plant infrastructure, which in turn led to lower than normal projected costs for decommissioning
497 and operations and maintenance.

498



499

500 **Fig. 6: LCOE estimates for Reykjanes and Vendenheim (US \$ per MWh per annum)**

501

502

503
504
505
506
507
508
509
510
511
512
513
514
515
516
517
518
519
520
521
522
523
524
525
526
527
528
529
530
531
532
533
534
535
536
537
538
539
540
541
542
543
544
545
546
547
548
549
550
551
552

3.3 Reflection on other cost components – social LCA and non-use value – and integration of alternative cost estimation approaches

The results from this study are based upon a cost methodology designed to estimate the economic and environmental implications of the DEEPEGS project. These are two dimensions in a sustainability analysis of a project’s cost consequences, but they omit social implications, which in many nations constitute one of the main barriers to geothermal utilisation (Barich et al., 2021). Social LCA (S-LCA) was first conducted in the 1990s but it has yet to be developed to the extent of LCA or ELCC approaches (Iofrida et al., 2018). In practice, where such studies are available, S-LCA assessments often involve merging or comparing outcomes in terms of social impacts with quantitative results from separate assessments, such as LCA or ELCC (Sala et al., 2015; Iofrida et al., 2018). Rather than representing a single technique, S-LCA is currently considered to be a set of methods focused on assessing the potential or real social impacts of projects in relation to human capital, human well-being, cultural heritage and social behaviour (Chhipi-Shrestha et al., 2014). Outcomes in terms of impacts are expressed using either qualitative or quantitative data, or both. In contrast to LCA and ELCC methods, S-LCA has added complexity due to the need to incorporate the underlying values and valuations of multiple stakeholder interests, including individuals, communities, governments, business and NGOs (Sala et al., 2015). This is especially complex in situations where different cultural elements, values and lifestyles apply, and these are dynamic factors across the lifecycle of the product or project. It is likely for these reasons that, as far as the authors are aware, no S-LCA studies have been conducted in the published literature in connection to geothermal power plants.

Alternative economic approaches to valuing socio-cultural impacts are emerging in a geothermal context, and these have been focused on the integration of an ecosystem services perspective. This links impacts to ecosystems and changes in human well-being to either monetary or non-monetary information, with the former comparable to the other costs and benefits of the project. Landsberg et al. (2011) conducted a review of the extent to which Environmental Impact Assessment data could be used to quantify the value of multiple, and often simultaneous, impacts to human well-being. The conclusion of the authors was that the approach had potential in terms of stimulating greater stakeholder integration into decision-making, with the increased likelihood of stakeholders retaining well-being benefits when developments took place.

In recent times, researchers have sought to apply non-market valuation techniques to elicit preferences and estimate the economic value of preserving geothermal landscapes. Such studies have focused on eliciting single value aggregate cost estimates for the preservation of geothermal landscapes, in so doing forming a cost estimate suitable for inclusion in a cost-benefit analysis for the project, and/or as a complement or addition to ELCC. Often the contingent valuation is applied to survey the economic value of sites where few or no people live or visit, or ever intend to visit. However, individuals will often still hold a preservation value for landscapes, and this is known as ‘non-use value’. Geothermal landscapes are a classic example of sites with likely non-use value due to their aesthetical rarity, diversity and distinctness. Despite controversy, largely centred around its reliance on constructed markets, the contingent valuation method is a survey-based technique that has been implemented to estimate willingness to pay for the preservation of geothermal areas. The first study by Thayer (1981) estimated the economic value of preserving a geothermal area in the Jemez Mountains of Santa Fe National Forest in New Mexico. More recently, the second and third contingent

553 valuation surveys were conducted to estimate the economic value of preserving the Eldvörp
554 and Hverahlíð geothermal fields in Iceland, the location of the former site is approximately ten
555 kilometres from the Reykjanes project (Cook et al., 2018). The studies provided survey
556 participants with real development scenarios in the form of likely geothermal power projects at
557 the study sites, together with a summary of their environmental and socio-cultural implications.
558 Total willingness to pay for preservation, aggregated across the affected Icelandic population,
559 was 2.10 and 1.77 billion ISK for Eldvörp and Hverahlíð, respectively. These values were found
560 to be equivalent to approximately 2% of the total construction costs of the Hellisheiði
561 Geothermal Power Plant, Iceland's largest and the world's third biggest in installed capacity.
562 No such studies have yet been conducted in the context of EGS or the development of a new
563 well on a site with an existing power plant, so these cost estimates should not necessarily be
564 considered indicative of the likely scale of the impacts at the DEEPEGS case study sites in
565 Reykjanes and Vendenheim.

566
567
568

569 3.4 Decision-making consequences of study

570

571 Outcomes from LCC or ELCC studies such as this one have considerable implications in terms
572 of supply chain management. Slightly more than 83% of the costs for the Reykjanes site related
573 to the well drilling phase, and a further 8% were projected to derive from the well stimulation.
574 Given Iceland's remote, island location in the North Atlantic, some shipping costs relating to
575 imported materials, such as the steel well casing, will be difficult to reduce. However,
576 experience gained from this inaugural, speculative project, will greatly assist in future
577 purchasing decisions and the determination of resource allocations, likely enabling knowledge
578 gained from this project to lower future ELCC outcomes for EGS ventures in Iceland, France
579 and elsewhere. In particular, there may be opportunities to lower the costs of operational
580 externalities in future projects, perhaps via the integration of proven carbon and hydrogen
581 sulphide capture technologies, such as those demonstrated in the CarbFix and SulFix projects
582 in Iceland (Karlsdottir et al., 2020; Kristjánsdóttir, 2014).

583

584 Some cost components in this study would also be of relevance to any cost benefit analysis
585 conducted for the Reykjanes or Vendenheim projects. Traditional forms of cost benefit analysis
586 seek to account for most cost components studied in this project, however, they are rarely, if
587 ever, extended to account for the economic value of environmental impacts. Thus, all costs in
588 this study relating to process-based 'internal' environmental externalities and those from
589 operations are normally omitted. Although relatively small in comparison to the total ELCC
590 estimate (especially so in the case of Vendenheim), these costs are non-negligible for Reykjanes
591 since they comprise more than 1% of the total. Moreover, if impacts to other ecosystem services,
592 such as non-use value, were also accounted for, then the likelihood increases that the economic
593 value of environmental and social impacts could have a decisive impact on the welfare gains or
594 losses of a project, as assessed in cost benefit analysis. These impacts could be considerable in
595 the context of projects introducing new production plant facilities, with potential visual and
596 aesthetic impacts over a wide geographical area. However, they may equally be less significant
597 when the production site is a former fossil fuel plant facility, as per Vendenheim.

598

599 Additionally, there are sometimes wider issues pertaining to the valuation of ecosystem service
600 impacts from developing geothermal resources, which ensure that not all costs and impacts
601 should be estimated using monetary information, as has already been briefly discussed in the
602 context of S-LCA. Although unlikely to apply to the case of either Reykjanes or Vendenheim,

603 environmental and social impacts often occur in relation to a wide range of stakeholders,
604 including indigenous communities. Where indigenous societies, such as the Kenyan Maasai or
605 New Zealand's Maori, hold deep and resonant values in relation to a geothermal resource,
606 leading to ecosystem services such as spiritual enrichment or community identity, these will
607 typically be formed on a collective rather than individual basis (Cook et al., 2019; Cook et al.,
608 2020). Although it is possible to estimate their value using either qualitative or quantitative data,
609 the use of a money metric to value individual preferences and willingness to pay is an
610 inappropriate means of estimating impacts to collectively formed non-material well-being. This,
611 in turn, means that it is not sufficient for decision-makers to simply be informed about a
612 geothermal power project's direct economic costs and benefits – for example, using cost benefit
613 analysis and/or an ELCC outcome – prior to determining whether to proceed with the venture.
614 Rather, advanced decision-support platforms need to be utilised, which embed and integrate
615 monetary and non-monetary information concerning the welfare implications of the power
616 project. This theme is explored in more detail in an earlier paper by Cook et al. (2019), in which
617 Multi Criteria Decision Analysis (MCDA) is advanced as a potentially suitable platform for
618 this endeavour. A review is conducted of the extent to which existing MCDA analyses of
619 geothermal power projects evaluate stakeholder interests and seek to account for an array of
620 values in decision-making, both monetary and socio-cultural. It is evident that a small number
621 of MCDA studies have ranked or prioritized variables (technical, economic, environmental and
622 social), including those by Borzoni et al., (2014); Jesus, (1997) and Polatidis et al. (2015).
623 However, other MCDA studies in the review by Cook et al. (2019) tended to have a limited
624 focus on only economic variables and were lacking in-depth stakeholder consultation.

625

626 3.5 Study limitations

627

628 Akin to the results from the LCA studies pertaining to the two sites, caution should be expressed
629 with regards to the ELCC estimates in this paper, given that they are made prior to
630 commencement of production from the new wells at both sites. Thus, various assumptions had
631 to be made, particularly those relating to power outputs, capacity factors and future variable
632 costs. The range of discount factors (2-8%) may also be insufficient to represent the risk
633 associated with deep EGS projects, which is a speculative technology in its infancy, and the
634 assumed duration of both projects of 30 years may be too conservative. In addition, although
635 ultimately intended to address four environmental impact categories – human health, ecosystem,
636 resource availability and climate change – the current ReCiPe tool only monetises externalities
637 linked to climate change. ReCiPe is thus an emerging approach, lacking data and not yet
638 sufficient in scope to account for the full spectrum of externalities from a geothermal power
639 project. In addition, certain costs, such as impacts to socio-cultural values are not covered by
640 the four categories of costs that the ReCiPe method seeks to address, and these were thus not
641 included within these studies. This has also been the case with other, widely popularised
642 approaches to capturing external costs, such as the European Commission's ExternE project
643 from the 1990s.

644

645

646

647 4. Conclusion

648

649 This study has provided the first estimates (non-comparable) of the environmental life cycle
650 costs of developing deep EGS at Reykjanes, Iceland and Vendenheim, France. Inclusive of
651 environmental externalities, the levelized cost of energy for the Reykjanes and Vendenheim
652 sites is 16.5 and 45.0 €/MWh/year, respectively. The outcomes from this study are of interest

653 and relevance to public procurement agencies, supply chain managers, purchasers and other
654 decision-makers, including EU funding agencies who call for such studies to be carried out.
655 Cost components linked to externalities, both ‘internal’, related to well drilling and stimulation
656 processes, and external, relating to operations after production commences are of particular
657 relevance to decision-making processes. Together, these contributed more than 2% of the total
658 environmental lifecycle costs in the case of the Reykjanes site. More work is needed to ensure
659 that the social costs of developing geothermal power projects are accounted for in ELCC studies,
660 perhaps involving the use of S-LCA alongside LCA and ELCC. Moreover, the necessity of
661 providing decision-makers with an array of monetary and non-monetary information
662 concerning the environmental and social implications of geothermal power projects
663 promulgates the need for advanced platforms to integrate such data. In this study, Multi-Criteria
664 Decision Analysis is briefly discussed as being one such tool, however, more research and
665 advancement is needed in order to ensure that such tools adequately reflect stakeholder interests
666 and impacts to multiple and diverse values linked to geothermal areas.
667

668

669 **Acknowledgements**

670 This paper has been subject to funding from the European Union’s Horizon 2020 research
671 programme in relation to the DEEPEGS project (grant no. 690771).
672

673

674 **References**

675

676 Ármannsson, H. (2018). An overview of carbon dioxide emissions from Icelandic geothermal
677 areas. *Applied Geochemistry*, 97, 11-18.
678

679

679 Barich, A., Stokłosa, A. W., Hildebrand, J., Eliasson, O., Medgyes, T., Quinonez, G., ... &
680 Fernandez, I. (2021). Social License to Operate in Geothermal Energy. *Energies*, 15(1), 139.
681

682

682 Barringer, H. P. (2003, May). A life cycle cost summary. In *International conference of*
683 *maintenance societies* (pp. 20-23).
684

685

685 Batumbya, B., Liu, J., Damien, W., & Lukawski, T. (2006). Life Cycle Assessment for Wind
686 Turbine. *APER, Milan*.
687

688

688 Bayer, P., Rybach, L., Blum, P., & Brauchler, R. (2013). Review on life cycle environmental
689 effects of geothermal power generation. *Renewable and Sustainable Energy Reviews*, 26, 446-
690 463.
691

692

692 Borzoni, M., Rizzi, F., & Frey, M. (2014). Geothermal power in Italy: A social multi-criteria
693 evaluation. *Renewable Energy*, 69, 60-73.
694

695

695 Buonocore, E., Vanoli, L., Carotenuto, A., & Ulgiati, S. (2015). Integrating life cycle
696 assessment and emergy synthesis for the evaluation of a dry steam geothermal power plant in
697 Italy. *Energy*, 86, 476-487.
698

699

699 Campbell, P. K., Beer, T., & Batten, D. (2011). Life cycle assessment of biodiesel production
700 from microalgae in ponds. *Bioresource technology*, 102(1), 50-56.
701

701

702 Chakrabarty, R., Roy, T., & Chaudhuri, K. S. (2017). A production: inventory model for
703 defective items with shortages incorporating inflation and time value of money. *International*
704 *Journal of Applied and Computational Mathematics*, 3(1), 195-212.

705

706 Chhipi-Shrestha, G. K., Hewage, K., & Sadiq, R. (2015). ‘Socializing’ sustainability: a critical
707 review on current development status of social life cycle impact assessment method. *Clean*
708 *Technologies and Environmental Policy*, 17(3), 579-596.

709

710 Chiasson, A. (2006). Life-cycle cost study of a geothermal heat pump system. *Geo-Heat Center,*
711 *Oregon Institute of Technology, Klamath Falls, OR.*

712

713 Ciambrone, D. F. (2018). *Environmental life cycle analysis*. CRC Press.

714

715 Clark, C., Sullivan, J., Harto, C., Han, J., & Wang, M. (2012, January). Life cycle
716 environmental impacts of geothermal systems. In *Proceedings, Thirty-Seventh Workshop on*
717 *Geothermal Reservoir Engineering Stanford University, Stanford, California.*

718

719 Colucci, V., Manfreda, G., Mendecka, B., Talluri, L., & Zuffi, C. (2021). LCA and Exergo-
720 Environmental Evaluation of a Combined Heat and Power Double-Flash Geothermal Power
721 Plant. *Sustainability*, 13(4), 1935.

722

723 Cook, D., Davíðsdóttir, B., & Kristófersson, D. M. (2016). Energy projects in Iceland–
724 Advancing the case for the use of economic valuation techniques to evaluate environmental
725 impacts. *Energy Policy*, 94, 104-113.

726

727 Cook, D., Davíðsdóttir, B., & Kristófersson, D. M. (2017). An ecosystem services perspective
728 for classifying and valuing the environmental impacts of geothermal power projects. *Energy*
729 *for Sustainable Development*, 40, 126-138.

730

731 Cook, D., Davíðsdóttir, B., & Kristófersson, D. M. (2018). Willingness to pay for the
732 preservation of geothermal areas in Iceland–The contingent valuation studies of Eldvörp and
733 Hverahlíð. *Renewable Energy*, 116, 97-108.

734

735 Cook, D., Fazeli, R. & Davíðsdóttir, B. (2019). A need for integrated valuation tools to support
736 decision-making processes – the case of cultural ecosystem services sourced from geothermal
737 areas. *Ecosystem Services*, 37, 100923.

738

739 Cook, D., Davíðsdóttir, B., & Malinauskaite, L. (2020). A cascade model and initial exploration
740 of co-production processes underpinning the ecosystem services of geothermal
741 areas. *Renewable Energy*, 161, 917-927.

742

743 Ebenhoch, R., Matha, D., Marathe, S., Munoz, P. C., & Molins, C. (2015). Comparative
744 levelized cost of energy analysis. *Energy Procedia*, 80, 108-122.

745

746 EIB. (2015). EIB Climate Strategy – Mobilising finance for the transition to a low-carbon and
747 climate-resilient economy. Available online:
748 https://www.eib.org/attachments/strategies/eib_climate_strategy_en.pdf (accessed 21st March
749 2022).

750

751 Estevan, H. & Schaefer, B. (2017). Life cycle costing – state of the art report. SPP Regions –
752 Regional Networks for Sustainable Procurement. Retrieved from:
753 http://www.sppregions.eu/fileadmin/user_upload/Life_Cycle_Costing_SoA_Report.pdf
754 (accessed 1st November 2018).
755

756 Frank, E. D., Sullivan, J. L., & Wang, M. Q. (2012). Life cycle analysis of geothermal power
757 generation with supercritical carbon dioxide. *Environmental Research Letters*, 7(3), 034030.
758

759 Friðleifsson, G. Ó, Elders, W.A., Zierenberg, R.A., Fowler, A.P.G., Weisenberger, T.B.
760 Mesfin, K.G., Sigurðsson, Ó., Níelsson, S., Einarsson, G., Óskarsson, F., Guðnason, E.Á.,
761 Tulinius, H., Hokstad, K., Benoit, G., Nono, F., Loggia, D., Parat, F., Sarah B. Cichy, S.B.,
762 Escobedo, D., Mainprice, D. The Iceland Deep Drilling Project at Reykjanes: Drilling into the
763 root zone of a black smoker analog, *Journal of Volcanology and Geothermal Research*,
764 Volume 391, 2020, 106435, ISSN 03770273, <https://doi.org/10.1016/j.jvolgeores.2018.08.013>
765

766 Friðleifsson, G. Ó., Elders, W. A., Zierenberg, R. A., Fowler, A. P., Weisenberger, T. B.,
767 Mesfin, K. G., ... & Guðnason, E. Á. (2020). The Iceland Deep Drilling Project at Reykjanes:
768 Drilling into the root zone of a black smoker analog. *Journal of Volcanology and Geothermal*
769 *Research*, 391, 106435.
770

771 Goh, B. H., & Sun, Y. (2016). The development of life-cycle costing for buildings. *Building*
772 *Research & Information*, 44(3), 319-333.
773

774 Gunnarsson, I., Aradóttir, E. S., Oelkers, E. H., Clark, D. E., Arnarson, M. Þ., Sigfússon, B., ...
775 & Gíslason, S. R. (2018). The rapid and cost-effective capture and subsurface mineral storage
776 of carbon and sulfur at the CarbFix2 site. *International Journal of Greenhouse Gas Control*, 79,
777 117-126.
778

779 Habibzadeh-Bigdarvish, O., Yu, X., Lei, G., Li, T., & Puppala, A. J. (2019). Life-Cycle cost-
780 benefit analysis of Bridge deck de-icing using geothermal heat pump system: A case study of
781 North Texas. *Sustainable cities and society*, 47, 101492.
782

783 Helgason, R., Cook, D., & Davíðsdóttir, B. (2020). An evaluation of the cost-competitiveness
784 of maritime fuels—a comparison of heavy fuel oil and methanol (renewable and natural gas) in
785 Iceland. *Sustainable Production and Consumption*, 23, 236-248.
786

787 Heralova, R. S. (2017). Life Cycle Costing as an Important Contribution to Feasibility Study in
788 Construction Projects. *Procedia Engineering*, 196, 565-570.
789

790 Huenges, E., & Ledru, P. (2010). *Geothermal Energy Systems: Exploration, Development, and*
791 *Utilization*, Wiley-VCH, Weinheim.
792

793 Hunkeler, D., Lichtenvort, K., & Rebitzer, G. (2008). *Environmental life cycle costing*. Crc
794 press.
795

796 Hutterer, G. W. (2020). Geothermal Power Generation in the World 2015-2020 Update Report.
797 Proceedings of the World Geothermal Congress 2020. Retrieved from:
798 <https://www.geothermal-energy.org/pdf/IGAstandard/WGC/2020/01017.pdf> (accessed 30
799 June 2020).
800

801 ICLEI. (nd). Local governments for sustainability. Retrieved from: <http://www.iclei.org/>
802 (accessed 7th October 2018).
803

804 Iofrida, N., De Luca, A. I., Strano, A., & Gulisano, G. (2018). Can social research paradigms
805 justify the diversity of approaches to social life cycle assessment?. *The International Journal*
806 *of Life Cycle Assessment*, 23(3), 464-480.
807

808 IRENA. (2018). Renewable Power Generation Costs in 2017. Retrieved from:
809 [https://www.irena.org//media/Files/IRENA/Agency/Publication/2018/Jan/IRENA_2017_Pow](https://www.irena.org//media/Files/IRENA/Agency/Publication/2018/Jan/IRENA_2017_Power_Costs_2018.pdf)
810 [er_Costs_2018.pdf](https://www.irena.org//media/Files/IRENA/Agency/Publication/2018/Jan/IRENA_2017_Power_Costs_2018.pdf) (accessed 9th December 2018).
811

812 Islam, S., Ponnambalam, S. G., & Lam, H. L. (2016). Review on life cycle inventory: methods,
813 examples and applications. *Journal of cleaner production*, 136, 266-278.
814

815 ISO (International Organization for the Standardization). (2017). ISO 15686-5:2017. Buildings
816 and constructed assets – service life planning – part 5: life-cycle costing. Retrieved from:
817 <https://www.iso.org/obp/ui/#iso:std:iso:15686:-5:en> (accessed 13th October 2018).
818

819 Jackson Jr, D. W., & Ostrom, L. L. (1980). Life cycle costing in industrial purchasing. *Journal*
820 *of Purchasing and Materials Management*, 16(4), 8-12.
821

822 Jaggi, C., Khanna, A., & Nidhi, N. (2016). Effects of inflation and time value of money on an
823 inventory system with deteriorating items and partially backlogged shortages. *International*
824 *Journal of Industrial Engineering Computations*, 7(2), 267-282.
825

826 Jesus, A. C. D. (1997). Environmental sustainability of geothermal development. *Energy*
827 *sources*, 19(1), 35-47.
828

829 Kambanou, M. L., & Sakao, T. (2020). Using life cycle costing (LCC) to select circular
830 measures: A discussion and practical approach. *Resources, Conservation and Recycling*, 155,
831 104650.
832

833 Karlsdottir, M. R., Heinonen, J., Palsson, H., & Palsson, O. P. (2020). Life cycle assessment of
834 a geothermal combined heat and power plant based on high temperature
835 utilization. *Geothermics*, 84, 101727.
836

837 Kerzner, H., & Kerzner, H. R. (2017). *Project management: a systems approach to planning,*
838 *scheduling, and controlling*. John Wiley & Sons.
839

840 Kristjánisdóttir, Helga. (2014). "The SulFix Procedure." In *Economics and Power-intensive*
841 *Industries*, pp. 59-66. Springer, Cham.
842

843 Landsberg, F., Ozment, S., Stickler, M., Henninger, N., Treweek, J., Venn, O., & Mock, G.
844 (2011). Ecosystem services review for impact assessment. World Business Council for
845 Sustainable Development and World Resources Institute.
846

847 Lazard. (2018). Lazard's Levelized Cost of Energy Analysis. Version 12.0. Lazard, Bermuda.
848 Retrieved from: [https://www.lazard.com/media/450784/lazards-levelized-cost-of-energy-](https://www.lazard.com/media/450784/lazards-levelized-cost-of-energy-version-120-vfinal.pdf)
849 [version-120-vfinal.pdf](https://www.lazard.com/media/450784/lazards-levelized-cost-of-energy-version-120-vfinal.pdf) (accessed 9th December 2018).
850

851 Lindholm, A., Laine, T. J., & Suomala, P. (2017). The potential of management accounting and
852 control in global operations: Profitability-driven service business development. *Journal of*
853 *Service Theory and Practice*, 27(2), 496-514.

854
855 Lindholm, A., Korhonen, T., Laine, T., & Suomala, P. (2018). Engaging the economic facts
856 and valuations underlying value for money in public procurement. *Public Money &*
857 *Management*, 1-10.

858
859 Luttenberger, A., & Luttenberger, L. R. (2017). Sustainable procurement and environmental
860 life-cycle costing in maritime transport. *WMU Journal of Maritime Affairs*, 16(2), 219-231.

861
862 Martínez-Corona, J. I., Gibon, T., Hertwich, E. G., & Parra-Saldívar, R. (2017). Hybrid life
863 cycle assessment of a geothermal plant: From physical to monetary inventory
864 accounting. *Journal of cleaner production*, 142, 2509-2523.

865
866 Menberg, K., Heberle, F., Bott, C., Brüggemann, D., & Bayer, P. (2021). Environmental
867 performance of a geothermal power plant using a hydrothermal resource in the Southern
868 German Molasse Basin. *Renewable Energy*, 167, 20-31.

869
870 Ney, R. A. & Schnoor, J. L. (2002). Greenhouse gas emission impacts of substituting
871 switchgrass for coal in electric generation: the Chariton valley biomass project. Centre for
872 Global and Regional Environmental Research, University of Iowa.

873
874 Ólafsson, K. B. (2016). Levelized Cost of Energy og virkjunarkostir til umfjöllunar í 3. Áfanga
875 rammaáætlunar. Powerpoint presentation retrieved from: [https://www.samorka.is/wp-](https://www.samorka.is/wp-content/uploads/2016/07/2016-LCOE-greiningarsk%C3%BDrsla-Samorku-uppf%C3%A6rlsa-11.-j%C3%BAI%C3%AD-og-til-prentunar-LOK.pdf)
876 [content/uploads/2016/07/2016-LCOE-greiningarsk%C3%BDrsla-Samorku-](https://www.samorka.is/wp-content/uploads/2016/07/2016-LCOE-greiningarsk%C3%BDrsla-Samorku-uppf%C3%A6rlsa-11.-j%C3%BAI%C3%AD-og-til-prentunar-LOK.pdf)
877 [uppf%C3%A6rlsa-11.-j%C3%BAI%C3%AD-og-til-prentunar-LOK.pdf](https://www.samorka.is/wp-content/uploads/2016/07/2016-LCOE-greiningarsk%C3%BDrsla-Samorku-uppf%C3%A6rlsa-11.-j%C3%BAI%C3%AD-og-til-prentunar-LOK.pdf) (accessed 30 June
878 2020).

879
880 Park, S., Langat, A., Lee, K., & Yoon, Y. (2021). Measuring the impact of risk on LCOE
881 (levelized cost of energy) in geothermal technology. *Geothermal Energy*, 9(1), 1-19.

882
883 Paulillo, A., Striolo, A., & Lettieri, P. (2019). The environmental impacts and the carbon
884 intensity of geothermal energy: A case study on the Hellisheiði plant. *Environment*
885 *international*, 133, 105226.

886
887 Polatidis, H., Haralambidou, K., & Haralambopoulos, D. (2015). Multi-criteria decision
888 analysis for geothermal energy: A comparison between the ELECTRE III and the
889 PROMETHEE II methods. *Energy Sources, Part B: Economics, Planning, and Policy*, 10(3),
890 241-249.

891
892 Pratiwi, A., Ravier, G., & Genter, A. (2018). Life-cycle climate-change impact assessment of
893 enhanced geothermal system plants in the Upper Rhine Valley. *Geothermics*, 75, 26-39.

894
895 Rebitzer, G., & Hunkeler, D. (2003). Life cycle costing in LCM: ambitions, opportunities, and
896 limitations.

897
898 Sala, S., Benini, L., Mancini, L., & Pant, R. (2015). Integrated assessment of environmental
899 impact of Europe in 2010: data sources and extrapolation strategies for calculating
900 normalisation factors. *The International Journal of Life Cycle Assessment*, 20(11), 1568-1585.

901
902 Sanjuan, B., Negrel, G., Le Lous, M., Poulmarch, E., Gal, F., & Damy, P. C. (2020, April).
903 Main geochemical characteristics of the deep geothermal brine at Vendenheim (Alsace, France)
904 with constraints on temperature and fluid circulation. In *World Geothermal Congress 2020*.
905 Available online:
906 <https://hal.archives-ouvertes.fr/hal-02335810/file/SanjuanetalPaperWGC2020-VF.pdf>
907 (accessed 21/03/2022).
908

909 Santoyo, E., Acevedo-Anicasio, A., Pérez-Zarate, D., & Guevara, M. (2018). Evaluation of
910 artificial neural networks and eddy covariance measurements for modelling the CO2 flux
911 dynamics in the Acoculco geothermal caldera (Mexico). *International Journal of*
912 *Environmental Science and Development*, 9, 298-302.
913

914 Shortall, R., Davidsdottir, B., & Axelsson, G. (2015). Geothermal energy for sustainable
915 development: A review of sustainability impacts and assessment frameworks. *Renewable and*
916 *sustainable energy reviews*, 44, 391-406.
917

918 Sigfússon, B., Arnarson, M. P., Snæbjörnsdóttir, S. Ó., Karlsdóttir, M. R., Aradóttir, E. S., &
919 Gunnarsson, I. (2018). Reducing emissions of carbon dioxide and hydrogen sulphide at
920 Hellisheidi power plant in 2014-2017 and the role of CarbFix in achieving the 2040 Iceland
921 climate goals. *Energy Procedia*, 146, 135-145.
922

923 Sigurjónsson, H. Æ., Cook, D., Davíðsdóttir, B., & Bogason, S. G. (2021). A life-cycle analysis
924 of deep enhanced geothermal systems—The case studies of Reykjanes, Iceland and Vendenheim,
925 France. *Renewable Energy*, 177, 1076-1086.
926

927 Spath, P. L., & Mann, M. K. (2000). *Life cycle assessment of a natural gas combined-cycle*
928 *power generation system* (pp. 1-56). Golden, CO: National Renewable Energy Laboratory.
929

930 Stark, J. (2015). Product lifecycle management. In *Product Lifecycle Management (Volume*
931 *1)* (pp. 1-29). Springer, Cham.
932

933 Sullivan, J. L., Clark, C. E., Yuan, L., Han, J., & Wang, M. (2012). *Life-cycle analysis results*
934 *for geothermal systems in comparison to other power systems: Part II* (No. ANL/ESD/11-12).
935 Argonne National Lab.(ANL), Argonne, IL (United States).
936

937 Swarr, T. E., Hunkeler, D., Klöpffer, W., Pesonen, H. L., Citroth, A., Brent, A. C., & Pagan, R.
938 (2011). Environmental life-cycle costing: a code of practice.
939

940 Thayer, M. A. (1981). Contingent valuation techniques for assessing environmental impacts:
941 further evidence. *Journal of Environmental Economics and Management*, 8(1), 27-44.
942

943 Tomasini-Montenegro, C., Santoyo-Castelazo, E., Gujba, H., Romero, R. J., & Santoyo, E.
944 (2017). Life cycle assessment of geothermal power generation technologies: An updated
945 review. *Applied Thermal Engineering*, 114, 1119-1136.
946

947 Tsang, A. H., Yeung, W. K., Jardine, A. K., & Leung, B. P. (2006). Data management for CBM
948 optimization. *Journal of quality in maintenance engineering*, 12(1), 37-51.
949

950 Van den Abeele, É. (2014). Integrating social and environmental dimensions in public
951 procurement: one small step for the internal market, one giant leap for the EU?. Working paper
952 2014.08, European Trade Union Institute.

953

954 Yilmaz, C. (2020). Life cycle cost assessment of a geothermal power assisted hydrogen energy
955 system. *Geothermics*, 83, 101737.

956